

1 **Strategies for Mitigating Impacts to AeroFauna from Offshore Wind Energy Development:**
2 **Available Evidence and Data Gaps**

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13 **Abstract**

14 Offshore wind energy (OSW) development, while a key strategy for reducing global reliance on fossil
15 fuels, nevertheless has environmental effects that should be mitigated. We reviewed the scientific
16 literature and gray literature to identify approaches for mitigating (e.g., avoiding, minimizing, or
17 compensating for) the effects of OSW development on birds and bats (aerofauna). The review included
18 studies from other industries where relevant, including terrestrial wind energy and the offshore oil and gas
19 industry. Of a total of 212 mitigation approaches from 233 source documents, 59% of proposed
20 approaches were not tested in the reviewed literature to assess effectiveness at mitigating anthropogenic
21 impacts to aerofauna. Of the mitigation approaches that were field tested or implemented, the reviewed
22 literature indicated evidence of their effectiveness in only about 36% of cases. Thus, there was no
23 evidence of effectiveness for 86% of the mitigation approaches identified in this literature review. For
24 birds, minimization approaches related to lighting (e.g., reducing artificial light, avoiding white and
25 steady-burning lights) were the most commonly tested and effective methods for reducing maladaptive
26 attraction and collisions. For bats, minimization via alteration of turbine operations (e.g., curtailment and
27 feathering of turbine blades) were most commonly shown to be effective. Minimization was the main
28 focus of this review but there is limited evidence of effectiveness for most approaches, and we suggest
29 implementation of dedicated testing to explore the effectiveness of commonly suggested and implemented
30 mitigation measures such as curtailment for birds. As such, avoidance of effects (via careful siting of
31 industrial activity and related measures to avoid effects to wildlife and their habitats) remains the best
32 available option for mitigation. To fully mitigate the effects of OSW development on aerofauna,
33 compensation and offset strategies should also be further explored.

34 **Keywords**

35 Mitigation, birds, bats, offshore wind energy development, minimization, avoidance, effectiveness

36 1. Introduction

37 OSW development is rapidly growing worldwide. While this growth supports the reduction of greenhouse
38 gas emissions, a critical step to minimizing the effects of climate change on wildlife, OSW development
39 also represents potential risks to wildlife (Peste et al., 2015). Wildlife may be affected with risk of direct
40 effects (e.g., injury/mortality via collisions) or indirect effects (e.g., altered movement patterns, behavior,
41 habitat; (Goodale and Stenhouse, 2016). Many bird and bat species around the world are in decline and
42 are protected at the state, provincial, federal, or international level (Dias et al., 2019; Frick et al., 2020;
43 Neate-Clegg et al., 2020), and the added risks posted by OSW development are important to consider
44 with this context. There are three main types of effects to aerofauna from OSW development: collisions,
45 behavioral change, and habitat-mediated change (Williams et al., 2024). Our scientific understanding of
46 the scale of these effects varies widely. Much of what we know about these pathways and effects come
47 from the offshore wind industry in Europe as well as the onshore wind industry in North America and
48 Europe; however, numerous gaps remain in our understanding (Williams et al., 2024).

49
50 Collisions that can lead to injury and mortality are well-documented for bats and avian migrants (e.g.,
51 songbirds, raptors) at onshore facilities (Allison et al., 2019). Indeed, concern over high mortality levels,
52 particularly for migratory bats, has led to a large focus on collision mitigation in the onshore wind
53 industry in North America (Adams et al., 2021). There are challenges to measuring collisions offshore, as
54 the key detection method onshore (carcass searches) is largely impossible in the offshore environment.
55 While there have been a few documented collisions of marine birds at offshore wind facilities in Europe
56 via direct observation and camera systems (Skov et al., 2018; Tjørnløv et al., 2023), this phenomenon
57 remains less understood for aerofauna in marine systems than in the terrestrial context. The stressor in the
58 case of collisions is primarily the long-term presence of structures (e.g., novel turbines and substations in
59 the marine environment), though increased risk of collision can also occur due to changes in behavior
60 (e.g., attraction) and habitat-mediated change (e.g., increased prey around turbines). Given the potential
61 for collisions to directly influence population size (as well as relative ease in detecting and measuring
62 effects, at least onshore), many wind energy mitigation strategies for aerofauna focus on collision
63 minimization.

64
65 Behavioral change in response to the presence and activities of offshore wind farms can take multiple
66 forms, including avoidance, attraction, and displacement from ideal distributions (Williams et al., 2024).
67 Behavioral change has been documented for a range of marine bird species at offshore wind facilities in
68 Europe, with the type and degree of avoidance depending on many factors including species, facility
69 location and design (Dierschke et al., 2016; Lamb et al., 2024). The mechanisms driving these behavioral
70 responses are likely varied and may be difficult to tease apart in some cases. Attraction may be due to
71 increased opportunities for perching and roosting (Dierschke et al., 2016), lighting on turbines and
72 substations (Rebke et al., 2019), or habitat-mediated change (e.g., increased prey; (Russell et al., 2014).
73 Artificial light, and specifically steady-burning white light, has been shown to negatively affect a wide
74 range of wildlife (Burt et al., 2023). Birds tend to be attracted to and disoriented by artificial light at night,
75 particularly during migration; this can lead to death via collisions or via exhaustion (as they circle the
76 light source for long periods; (Montevecchi, 2006). Though less commonly reported, artificial light may
77 also cause avoidance in seabirds (Syposz et al., 2021). Artificial lighting is widely known to attract bats,
78 in part due to the attraction of insects on which bats feed (Stone et al., 2015). Although at least one study
79 suggests that lighting of onshore turbines does not affect bat mortality rates (Johnson et al., 2003),

80 lighting on offshore wind turbines, vessels, and substations has been identified as a potential risk to
81 aerofauna (Williams et al., 2024).

82
83 Avoidance of individual turbines within the wind farm footprint (meso-avoidance), and avoidance of the
84 entire offshore wind facility (macro-avoidance) with resultant changes in distributions (e.g.,
85 displacement) may relate to the presence of the offshore wind facility infrastructure itself, or associated
86 activities, such as vessel traffic (Cook et al., 2018). Behavioral response may also be a result of sound
87 during construction or operations or electromagnetic fields and other sensory changes (e.g., heat or
88 vibration), most of which occur underwater at OSW facilities. The effects of sound have been
89 documented in marine mammals (Madsen et al., 2006; Thomsen et al., 2006), but this is not well studied
90 in aerofauna.

91
92 Finally, habitat-mediated effects to aerofauna include changes in the physical environment and/or
93 resource availability. There are multiple stressors that may influence aerofauna (particularly marine birds)
94 around offshore wind facilities via changes in water quality, community alteration, introduction of
95 invasive species, scouring, and bottom disturbance. In addition, the presence of the structures themselves
96 may lead to changes in resources (e.g., artificial reef effect; (Degraer et al., 2020). All of these may lead
97 to changes in the physical environment that individuals experience, and may also influence the
98 distribution, abundance, and structure of prey communities. These habitat-mediated effects are difficult to
99 measure and generally remain poorly understood for aerofauna.

100
101 While we can gain understanding of these effects from areas where OSW development is a more
102 established industry (e.g., Europe), there remains a lot of uncertainty in potential effects and the
103 mechanisms driving these effects. In the face of this uncertainty, it is important to consider how best to
104 mitigate these effects. Mitigation of effects such as those posed to aerofauna by the OSW industry is
105 guided by a well-established hierarchy. This mitigation hierarchy has three main components: 1)
106 avoidance, 2) minimization, and 3) compensation. Given this framework, ‘mitigation’ hereafter is used to
107 refer to all steps in the mitigation hierarchy.

108
109 Avoidance is a first priority at the top of the hierarchy; the best strategy for dealing with anthropogenic
110 impacts is to avoid those impacts altogether. Subsequent steps occur in descending order of priority such
111 that effects that cannot be avoided should be minimized, then compensated for (Council on
112 Environmental Quality, 2020). From the perspective of mitigating the effects of OSW development on
113 aerofauna, avoidance primarily occurs during planning processes via the siting of projects away from high
114 risk areas (Arnett and May, 2016). Minimization primarily occurs during pre-construction and operations,
115 and can include design decisions (e.g., micro-siting, turbine height, lighting, addition of perching
116 deterrents) as well as operational parameters (e.g., cut-in speeds at which turbines begin generating
117 electricity). The final step in the mitigation hierarchy is compensation or offsets, whereby activities are
118 employed to improve and enhance impacted populations for which effects cannot be fully avoided or
119 minimized. Compensation can occur either on- or off-site and may include options such as habitat
120 creation (e.g., new breeding, roosting, wintering sites) and improvement of existing habitat via invasive
121 species removal or predator control (Arnett and May, 2016). Compensation aims to ensure that negative
122 environmental impacts that cannot be avoided or minimized can be moderated by environmental gains to

123 produce net neutral or net positive outcomes for affected populations and ecosystems (Kiesecker et al.,
124 2010).

125
126 While various mitigation approaches have been proposed for the offshore wind industry, often drawing
127 upon efforts from other industries such as terrestrial wind and offshore oil and gas, many methods remain
128 untested in the offshore wind context. Given the harsh and inaccessible nature of the offshore
129 environment, it can be both challenging and expensive to test mitigation strategies. Moreover, high levels
130 of environmental variability in the marine environment, often coupled with low sample sizes for studies
131 that are trying to measure change in these systems, create challenges in achieving adequate statistical
132 power to provide evidence of effectiveness (Maclean et al., 2013; Regional Synthesis Workgroup of the
133 Environmental Technical Working Group, 2023). Additionally, wildlife species and their habitats may be
134 affected by a range of effects and stressors, as noted above, and the sensitivity and vulnerability of species
135 varies in relation to a variety of factors, both internal (e.g., demographics, life history), and external (e.g.,
136 facility layout, location). As a result, many mitigation measures may or may not be effective, and in some
137 cases have already been implemented without the necessary evidence that they are appropriate and
138 effective approaches in the offshore context. This can lead to wasted resources and potential false
139 assurance that effects of OSW development are being appropriately mitigated.

140
141 The goal of this review is to assess the state of knowledge for mitigation for aerofauna in relation to OSW
142 development. This includes an inventory of the different mitigations that have been proposed for use for
143 OSW development across jurisdictions. Our focus is on mitigation for aspects of projects in the marine
144 environment (e.g., project footprint, cable and transit routes) as opposed to onshore aspects (e.g.,
145 transmission interconnection), recognizing that some mitigation measures discussed in this review may
146 also apply to onshore components of projects. We aim to summarize the weight of evidence supporting
147 various mitigation approaches and identify key knowledge gaps in our understanding, along with
148 potential criteria for prioritizing future studies to examine the effectiveness of mitigation options.

149

150 **2. Materials and Methods**

151 To assess the state of knowledge of OSW development mitigation for aerofauna, we utilized an existing
152 database of mitigation approaches and conducted an additional literature review to update the information
153 within.

154 **2.1 Mitigation Practices Database**

155 In 2018-2019, the New York State Energy Research and Development Authority funded a desktop review
156 to aggregate and assess practices for minimization and avoidance of impacts to fisheries and offshore
157 species (i.e., birds, bats, marine mammals, sea turtles, fish, and benthos) associated with all phases of
158 OSW development. The database developed from this literature review of existing mitigation approaches,
159 initially called the Mitigation and Monitoring Tool, was made publicly available online and in 2023 was
160 moved to a new location and renamed the Mitigation Practices Database (MPD) Tool
161 (<https://www.nyetwg.com/mpd-tool>). The sources included in the database for aerofauna (n=109)
162 included peer-reviewed and gray literature (e.g. agency reports, environmental assessments, technical
163 guidance documents) that were identified via Google Scholar (search terms: offshore wind + birds +

164 mitigation; offshore wind + bats + mitigation) and via an internal Biodiversity Research Institute
165 Mendeley database (search terms: mitigation; mitigate; minimize). Sources identified during this literature
166 review process cited other papers that were also included in the database where appropriate.

167
168 The structure of the database included information on the taxonomic group for which the mitigation was
169 proposed or tested, the types of stressors and effects that were addressed, the phase of development, and
170 evidence of effectiveness for the mitigation approach, among other information (Appendix A; Table S1).
171 This same structure was used for the subsequent update to the literature review conducted in 2024.
172 Effectiveness, in this context, was defined as the degree to which a mitigation approach was successful in
173 producing the desired result, which assumes scientific rigor in data collection for such a conclusion.

174 **2.2 Update to Literature Review**

175 An additional literature review was conducted to update information on mitigation approaches for
176 aerofauna from 2019-2023. A series of literature searches were conducted (focusing primarily on sources
177 from 2019-2023) using Google Scholar (search terms: bird+wind+mitigation; bat+wind+mitigat*;
178 bat+wind+mitigat*+offshore; bird*+wind+mitigat*+offshore) and the Tethys Knowledge Base (Filters:
179 Wind energy content + birds or bats; search term “mitigation”). The abstract or summary of each source
180 was reviewed for initial relevance. If there was no mention of mitigation, avoidance, minimization,
181 compensation, or offsetting, sources were deemed not relevant and were not reviewed further. The
182 remaining sources were reviewed, and information was extracted following the structure of the MPD Tool
183 database, above. This initial filtering resulted in the identification of 129 additional sources, of which 78
184 were included in the literature review (the remaining 51 were deemed not relevant following a more in-
185 depth review of the entire document).

186

187 **2.3 Data Synthesis**

188 Additional sources from 2019-2023 were synthesized with the original mitigation strategies included in
189 the MPD Tool database based on commonalities. For example, if a source from the updated literature
190 review discussed the mitigation strategy ‘paint turbine blades’ and the same approach was already
191 included in the MPD Tool, then data from the new source was synthesized with the original entry for that
192 mitigation approach. However, if a new source identified a strategy not in the MPD Tool, it was added as
193 a new mitigation approach.

194

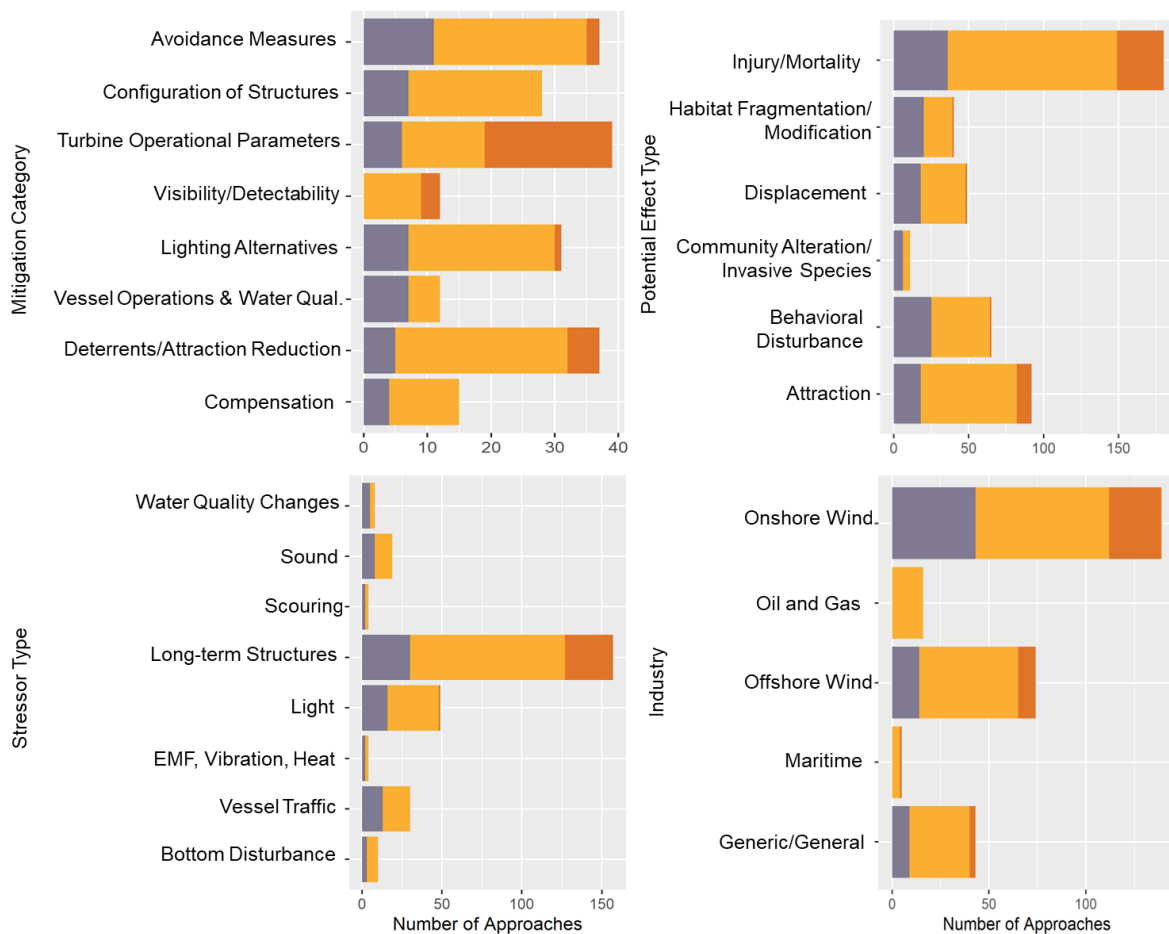
195 When examining the degree to which the various mitigation strategies have been tested for evidence of
196 effectiveness, we summarized by source document rather than by mitigation approach, as an individual
197 mitigation measure may have been suggested by multiple sources with various degrees of testing and
198 evidence of effectiveness. For the various mitigation approaches for aerofauna, we summarized by
199 mitigation category including the state of knowledge, level of testing and evidence of effectiveness, and
200 key knowledge gaps. The order of mitigation approaches discussed below generally follows the
201 mitigation hierarchy (avoid, minimize, compensate).

202

203 3. Results

204 3.1 Overall Summary

205 A broad range of mitigation approaches were identified from the existing literature, including 212
 206 approaches from 183 sources (Appendix B-C), focusing on a variety of stressors, potential effects, and
 207 industries. There were 76 sources from the literature review update, of which 28 were incorporated into
 208 existing approaches in the MPD Tool, while 48 identified new approaches. Mitigation approaches fell into
 209 8 general categories (Figure 1): 1) avoidance measures (e.g., wind farm siting and seasonal
 210 restrictions/limitations on activities, other activity limitations), 2) configuration of structures (e.g., wind
 211 farm layout, turbine size), 3) turbine operation parameters (e.g., curtailment, shutdown, feathering), 4)
 212 lighting to reduce attraction, 5) visibility and detection enhancement, 6) vessel operations and water
 213 quality, 7) deterrents and other attraction reduction measures, and 8) compensation/offsets.



214 **Figure 1.** Number of mitigation approaches by general category (top left), potential effect type (top right), stressor
 215 (bottom left), and industry (bottom right). Mitigation approaches were directed towards birds (yellow), bats
 216 (orange), or both taxa (gray). Mitigation categories are mutually exclusive; however, each mitigation approach could
 217 be listed under multiple stressors, effects, and industries.
 218
 219

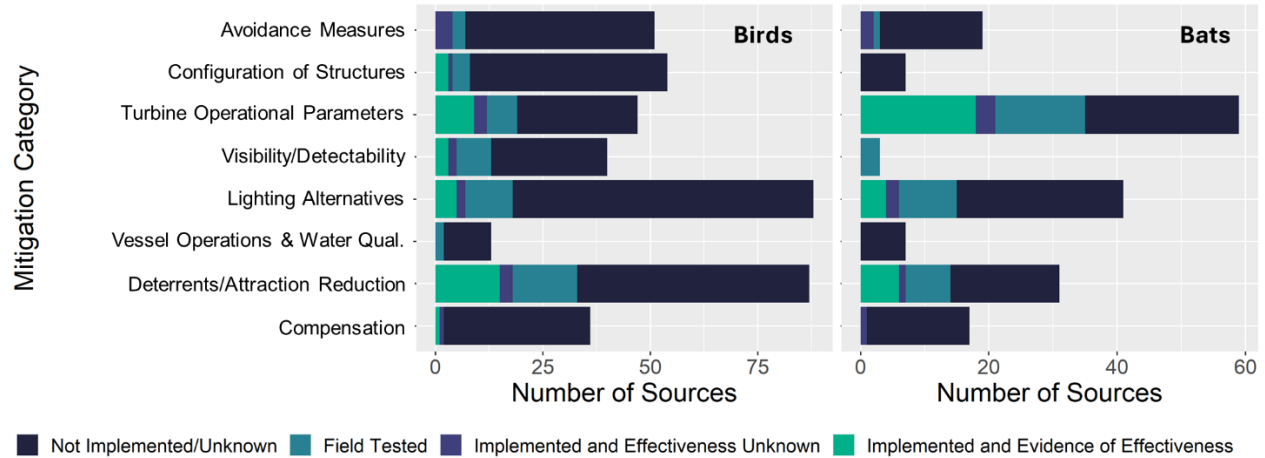
220 Across all categories, there were a wider variety of mitigation approaches focused on birds (n=180) than
221 bats (n=78), of which some were applicable to both (n=47). A few approaches (about 15% of the total)
222 were specific to marine birds or nocturnal avian migrants, but most were relatively general. Most
223 mitigation approaches fell into the minimization category of the mitigation hierarchy (n=178), followed
224 by avoidance (n=26), and lastly compensate/offset (n=14). Most mitigation approaches came at least in
225 part from the onshore wind industry (66%), with about 45% of mitigation approaches focused on
226 offshore/marine environments, including offshore oil and gas, offshore wind, and other maritime
227 industries (Figure 1); the remainder were “general” mitigation approaches that were not suggested in
228 relation to a specific type of industrial development, including studies focused on the agricultural sector.

229
230 Most mitigation approaches (74%) focused on the presence of long-term structures as the primary stressor
231 type, followed by lighting (23%) and changes in vessel traffic (14%). This is unsurprising given the
232 identification of long-term structures as the main source of OSW’s potential effects to aerofauna
233 (Williams et al., 2024). Most mitigation approaches focused on mitigating injury/mortality (85%) or
234 behavioral disturbance (31%; Figure 1). Within approaches focused on injury/mortality, 51% also focused
235 on attraction. The widest variety of suggested approaches was in the category of lighting and visibility
236 enhancement (n=43), most of which were focused on birds (74%). The turbine operation mitigations
237 included the greatest number of bat-related approaches (Figure 1), mostly variations of curtailment, which
238 is a key mitigation tool used for bats at onshore wind facilities (Adams et al., 2021; Whitby et al., 2021).

239
240 Almost 40% of mitigation approaches have been implemented or field tested (Figure 2), regardless of
241 category or taxonomic focus. An additional 19% had clearly not been implemented or tested. For the
242 remaining 41%, it was unclear if the approaches had been implemented or tested. It is likely that
243 implementing and testing of certain approaches were not explored due to feasibility, cost, or other
244 limiting factors. This means that 66% of approaches had likely not been tested to assess whether they
245 were effective. Of those mitigation approaches that were field tested or implemented, evidence of their
246 effectiveness was found in about 36% of cases (or 14% of all mitigation approaches in the database).
247 Thus, there was no evidence of effectiveness for 86% of the mitigation approaches identified in this
248 literature review.

249
250 The mitigation approaches for which there was evidence of effectiveness were primarily focused on birds,
251 or on both birds and bats (74%). However, a higher *ratio* of mitigation approaches that had been tested for
252 bats were shown to be effective (primarily approaches relating to curtailment). Unsurprisingly, given the
253 difficulties in testing mitigation approaches in the offshore environment, a slightly lower percentage of
254 offshore-focused methods had been field tested or implemented (about 36%) as opposed to 41-50% of
255 approaches targeted towards onshore, onshore/offshore, or generic situations. Where mitigation
256 approaches had been tested, it was typically at a small scale due to logistical difficulties, which led to
257 small sample sizes and other issues that may make it difficult to scale up results to understand the likely
258 influence of mitigation measures that are implemented more broadly.

259
260



261 ■ Not Implemented/Unknown ■ Field Tested ■ Implemented and Effectiveness Unknown ■ Implemented and Evidence of Effectiveness
 262 **Figure 2.** Number of sources by category for birds (left) and bats (right) for various mitigation
 263 approaches along with the level of testing and evidence for effectiveness (dark blue = mitigation approach
 264 was not implemented/tested in the source document or it was unclear whether it had been implemented;
 265 purple = mitigation approach had been field tested in the source document, but not implemented in a real-
 266 world situation; light blue = mitigation approach was implemented in a real-world situation, but it was
 267 either not tested to determine its effectiveness, or it was shown not to be effective; and green = mitigation
 268 approach was implemented and shown to be effective). Mitigation categories are mutually exclusive;
 269 however, if the same mitigation approach is discussed in multiple source documents within a category, it
 270 is represented multiple times in this figure (and may be classified into different effectiveness types, if the
 271 studies had differing results).

272
 273
 274 The most tested or implemented bird mitigation approaches were focused on minimization (i.e.,
 275 deterrents) and attraction reduction measures. Other minimization measures had high rates of testing,
 276 including lighting and visibility enhancement and turbine operation parameters. Much of the testing of
 277 these types of mitigation approaches occurred in relation to the onshore wind industry. For bat mitigation,
 278 turbine operation parameters (e.g., curtailment) had the highest testing rate and evidence of effectiveness,
 279 though lighting and visibility enhancement, and deterrence/attraction reduction, also exhibited high rates
 280 of testing and effectiveness. In contrast, mitigation methods focused on the configuration of wind farm
 281 structures, siting, seasonality, and compensation, were not well tested in the reviewed literature.

282 3.2 Avoidance Measures

283 Per the mitigation hierarchy, avoidance of impacts is the most effective and highest-priority strategy for
 284 mitigating anthropogenic impacts to wildlife and ecosystems (Council on Environmental Quality, 2020;
 285 Croll et al., 2022). Thirty-seven different mitigation approaches were suggested in 43 reviewed literature
 286 sources to avoid impacts to aerofauna by 1) siting development activities away from important
 287 populations or habitats (n=13 approaches), 2) conducting development activities outside of important
 288 habitat use periods (e.g., seasonal restrictions; n=5 approaches), 3) limiting activities (for example,
 289 phasing construction activities to avoid higher-risk time periods; n=15), and 4) communication and
 290 engagement with stakeholders (n=4). Approaches varied in their specifics, but many were very similar in
 291 concept. Surprisingly, given that siting is generally considered to be the best available mitigation

292 approach for avoiding risk to birds (Williams et al., 2024), the literature included some kind of
293 preliminary or desktop-based assessment of their effectiveness for only three of these proposed
294 approaches, and only included one document where an approach was actually implemented and showed
295 evidence of effectiveness.

296 *3.2.1 Siting and Spatial Avoidance*

297 Siting and other spatial-based avoidance approaches (n=15) included macro-siting (e.g., siting the entire
298 wind farm away from important habitat use areas) and micro-siting (e.g., placing individual turbines and
299 other infrastructure within a project footprint in such a way as to avoid key use areas; (Allison et al.,
300 2019). Important habitat use areas were defined in the literature in a variety of ways, including known bat
301 maternity roosts and hibernacula (American Wind Wildlife Institute, 2019; Reusch et al., 2022), avian
302 breeding colonies and nesting areas (Defingou et al., 2019; May et al., 2015; Simmons et al., 2020),
303 designated Important Bird Areas and other areas of high bird usage, e.g., with known natural
304 concentrations of birds (Drewitt and Langston, 2006; Langston and Pullan, 2003; Ldeke, 2019), known
305 feeding, staging, or loafing locations (Defingou et al., 2019; Ldeke, 2019; Thaxter et al., 2019),
306 migratory and commuting pathways (Defingou et al., 2019; Langston and Pullan, 2003; Ldeke, 2019;
307 Mockrin and Gravenmier, 2012), areas of high prey density (Watson et al., 2018), shallow waters and
308 waters in proximity to shore (Goodale and Milman, 2020; Lagerveld et al., 2023; Williams et al., 2015),
309 sensitive habitats (Mockrin and Gravenmier, 2012; Rydell et al., 2012), and undisturbed areas (Eshleman
310 and Elmore, 2013; Mockrin and Gravenmier, 2012). Several literature sources suggested using specific
311 types of data to help define these important habitat use areas, such as collision risk models, bat acoustic
312 activity data, weather data, and weather radar data on aerofauna migration (Cohen et al., 2022; Gaultier et
313 al., 2023; Murgatroyd et al., 2021). In addition to wind farm and turbine siting, other spatial avoidance
314 approaches included the implementation of buffers and setbacks in the choice of location or for particular
315 activities (Eshleman and Elmore, 2013; Mockrin and Gravenmier, 2012; Rydell et al., 2012).

316
317 While siting-related mitigation strategies are almost entirely applicable prior to construction of wind
318 farms, at least one review paper also suggested that micro-siting can be an important avoidance strategy
319 for repowering existing wind farms, e.g., via revision to wind farm layouts or decommissioning of
320 turbines that proved to be particularly problematic from the standpoint of wildlife effects (Marques et al.,
321 2014). However, this type of micrositing must be informed by data such as collision rates at individual
322 turbines, which are generally more accessible for terrestrial wind energy projects than offshore wind
323 facilities.

324
325 Most jurisdictions have implemented some type of spatial planning to inform siting projects, with at least
326 some consideration of wildlife-related impacts as part of that process (including other taxa as well as
327 aerofauna; for example, see (Farmer et al., 2023; Randall et al., 2022). Similar spatial planning efforts to
328 avoid important wildlife areas have also occurred for other industries such as terrestrial wind energy
329 development (Liang et al., 2023). Despite the common acceptance of avoidance as the most effective way
330 to mitigate impacts, relatively few studies in our database actually tested this assumption. Several studies
331 used different types of avian movement data to inform model-based assessments of risk, however; for
332 example, (Goodale et al., 2019) conducted a model-based analysis of Northern Gannet (*Morus bassanus*)
333 movements, and found that siting offshore wind projects away from shallow coastal waters would reduce
334 gannet exposure to development activities by avoiding key habitats. Likewise, (Cohen et al., 2022) used

335 weather radar data to assess putative collision risk based on stopover densities of migrants in the Great
336 Lakes region of North America, and found that siting projects >20 km from coastlines would help to
337 avoid this risk. (Murgatroyd et al., 2021) modeled movement data for Verraux's Eagle (*Aquila verreauxii*)
338 and found that a combination of nesting areas and topographic features could be used to predict collision
339 risk across the landscape, and thus inform siting decisions.

340 *3.2.2 Temporal Avoidance*

341 Several approaches focus on temporal, rather than spatial, separation (n=7). This involves scheduling
342 disruptive activities to avoid sensitive life history periods and/or times of high habitat use. This includes
343 scheduling construction activities to avoid the avian breeding season and other times of high bird or bat
344 abundance, such as migratory or wintering periods, depending on the taxa and location (Cook et al., 2011;
345 English et al., 2017; Eshleman and Elmore, 2013; Langston, 2013; May et al., 2015; Mockrin and
346 Gravenmier, 2012; Vaissière et al., 2014; World Bank Group, 2015). Additionally, avoiding scheduling
347 vessel and helicopter activity, as well as other maintenance activities, outside of sensitive life history
348 periods and times of peak habitat use at the development site was recommended (Gartman et al., 2016;
349 U.S. Fish and Wildlife Service (USFWS), 2021).

350

351 Seasonal limitations on construction activities to mitigate impacts on marine birds have been
352 implemented in multiple locations in the UK, with construction limited to March-October at the Burbo
353 Bank offshore wind farm to protect wintering birds, and restricted from 16 December-March at Rhyl Flats
354 to protect Common Scoter (English et al., 2017).

355 *3.2.3 Limiting Activities*

356 In addition to season-specific limitations, there are other types of activity limits that could be
357 implemented for mitigation that relate to the spatial and temporal overlap of development activities and
358 habitat for aerofauna. These include sequencing and phasing turbine installation and other construction
359 activities (English et al., 2017), limiting construction of artificial breeding structures in proximity to
360 development (Everaert and Stienen, 2007), and colocating equipment across projects (Drewitt and
361 Langston, 2008). Of the 14 strategies identified, two were implemented, four were not implemented, and
362 for seven strategies implementation status was unknown. The two implemented strategies were for
363 offshore wind in the UK. In one case, turbine installation progressed from north to south over the course
364 of a season to minimize the effect on a nearby Little Tern (*Sternula albifrons*) colony (English et al.,
365 2017); in the other example, a phased development approach was applied, with the second phase only
366 permitted once the developer demonstrated no construction-related change to Red-throated Loon (*Gavia*
367 *stellata*) habitat that would compromise identified conservation objectives (English et al., 2017). Despite
368 these examples, it is typically quite difficult to collect data on effectiveness for these types of mitigation
369 strategies (as with siting and seasonality approaches above).

370 *3.2.4 Engagement and Communication*

371 While not exclusive to siting, engagement and communication with stakeholders was identified as a
372 mitigation strategy in five reviewed papers, though none were tested for effectiveness. These
373 recommendations typically related to early and standardized stakeholder and community engagement.
374 Early interaction with state and federal agencies was also described, particularly in regard to siting, as a

375 method of minimizing the potential for impacts (U.S. Fish and Wildlife Service (USFWS), 2021). None
376 of these strategies were tested for effectiveness within the reviewed literature, likely as the majority of
377 these strategies were documented in guidance papers intended to provide planning assistance.
378

379 **3.3 Minimization Approaches**

380 *3.3.1 Configuration of Structures*

381 The configuration of offshore wind farms (and other anthropogenic infrastructure) has the potential to
382 influence how aerofauna perceive these structures and therefore how they respond to and interact with
383 them. Decisions on configuration occur during planning stages but often after macro-siting has already
384 occurred (e.g., the general location of the facility has been determined). Mitigation strategies (n=26
385 topics) included number of turbines, turbine layout (e.g., clustering, spacing), and turbine size (overall
386 height, size of “air gap” below the lower edge of rotor-swept zone), as well as minimizing the use of
387 wires that volant animals may be unable to see. Turbine layout considerations included placing turbines
388 within an array or clustered, and how much space should be between them. Considerations related to
389 clustering included migration corridors (Cook et al., 2011), leaving larger gaps between rows in the
390 direction of prevailing winds (Tulp et al., 1999), and identifying daily flight paths to foraging and
391 breeding sites to ensure safe passage through or around the wind farm (Vaissière et al., 2014).
392

393 The relationship between turbine size and mortality risk was variable, in part because “size” represented a
394 catchall term used to describe multiple turbine characteristics, including overall height, distance between
395 the surface of the water and the turbine blade (i.e., air gap), and rotor diameter (Garvin et al., 2024).
396 Increasing the air gap has been suggested to reduce mortality for marine birds that largely fly close to the
397 water’s surface. There have also been suggestions that increasing turbine height, and therefore reducing
398 the total number of turbines needed to produce the same amount of electricity, can help reduce collision
399 mortality. However, the ideal height is likely species and context specific. For example, multiple studies
400 found that increasing turbine height helped to reduce raptor mortalities at onshore wind farms (Arnett and
401 May 2016), including for Golden Eagles (*Aquila chrysaetos*; CEC 2002). However, the number of
402 neotropical migrant mortalities was two orders of magnitude higher for 300m turbines compared with
403 90m towers (Crawford and Engstrom, 2001). A recent study found rotor diameter was positively
404 correlated with bird mortalities but not with bat mortality rates (Garvin et al., 2024). The effectiveness of
405 this type of measure relates directly to flight height of species, and thus may be difficult to optimize for
406 multiple species and groups. There was little testing or evidence in other contexts, including for offshore
407 wind development. Because collision risk models are commonly used for the offshore wind industry,
408 model exercises could be performed to optimize height to reduce collision risk for particular species;
409 however, these models remained unvalidated in most cases.
410

411 *3.3.2 Turbine Operation Parameters*

412 Aerofauna collisions at onshore wind facilities (Schuster et al., 2015) have led this industry to focus on
413 mechanisms to reduce this risk. A primary focus to date has been on altering turbine operational
414 parameters and implementing shutdowns. Potential mitigation approaches related to turbine operational

415 parameters include 1) curtailment (n=26 approaches), 2) temporary shutdown (n=7 approaches), and 3)
416 feathering (n=5 approaches). Curtailment involves raising the threshold for wind speed (also called cut-in
417 speed) at which turbines begin to generate power. Below the cut-in speed, turbine blades spin but at much
418 slower speeds, if at all (blades can be “feathered,” or pitched to catch as little wind as possible; (Adams et
419 al., 2021). Because bat activity is typically higher at lower wind speeds, raising the cut-in speed for
420 turbines to start generating electricity ensures that there are fewer opportunities for negative interactions.
421 Other related turbine shutdown approaches have also been suggested or implemented.

422
423 While one strategy was to implement turbine curtailment consistently at night (termed blanket
424 curtailment; (Allison et al., 2019), this approach could lead to loss of energy production (Thurber et al.,
425 2023). As such there were various possible approaches for implementing more targeted curtailment
426 during periods of higher activity for aerofauna based on environmental conditions (e.g., a combination of
427 low wind speeds and higher temperatures; see (American Wind Wildlife Institute, 2019; Arnett and May,
428 2016; Sinclair and DeGeorge, 2016; Whitby et al., 2021), high risk periods of the year (e.g., migration;
429 (Hill et al., 2014; Smallwood and Bell, 2020), and/or the use of tools and technologies to detect bats or
430 birds within the vicinity of the turbine (e.g., radar, human observers, acoustic detectors) to trigger
431 curtailment, often termed ‘smart curtailment’. Various curtailment strategies were primarily field tested
432 and implemented in the onshore wind industry, both due to the longer history of this industry and relative
433 ease of detecting mortalities (via carcass searches) relative to offshore wind development.

434
435 There was strong evidence in the literature that curtailment was an effective mitigation strategy for bats in
436 the onshore wind context. Multiple review papers included in the database suggested that across studies, a
437 50% reduction in bat mortality could be achieved by implementing cut-in speeds 1.5-3.0 m/s above
438 normal (Garman et al. 2016). For example, a meta-analysis of bat curtailment studies in the Midwest and
439 Mid-Atlantic regions of the United States found that a 5.0 m/s cut-in speed could reduce total bat
440 mortalities across facilities by an average of 62% (Whitby et al., 2021). However, smart curtailment
441 approaches had not yet been shown to be effective for many bat species (Hayes et al., 2023), and this was
442 thus an area of active research in recent studies.

443
444 Given the effectiveness of curtailment for protecting bats, it was an oft-suggested mitigation approach for
445 birds at wind farms as well. However, there is currently little or no evidence that curtailment reduces
446 mortalities for birds (Smallwood and Bell, 2020). In general, the only studies with evidence of
447 effectiveness utilized human observers to implement smart curtailment with a focus on raptors (Ferrer et
448 al. 2022, Bennun et al. 2021). At least one automated detection technology to inform smart curtailment
449 for eagles has also been tested (McClure et al. 2018); while it was effective at detecting and classifying
450 eagles, no data were available in our review to determine its effectiveness at actually reducing collisions.
451 Evidence of the effectiveness of curtailment for other groups of birds was also generally lacking.
452 Curtailment was mentioned in the literature as a potential mitigation strategy in the offshore environment
453 for aerofauna, but this mitigation strategy remained untested in this context.

454
455 Similar to curtailment strategies, turbine shutdown mitigation strategies generally included temporary
456 shutdown of turbines during periods of high activity. Rather than partially basing these shutdowns on
457 wind speeds, periods of high activity were determined based on timing (e.g., seasonal) or other
458 environmental conditions or monitoring data (e.g., active monitoring or model predictions). These

459 strategies came primarily from the onshore industry, and in most cases lacked evidence of effectiveness,
460 though a review paper (Cook et al. 2011) suggested evidence of effectiveness for bats and Marbled
461 Murrelets (*Brachyramphus marmoratus*).

462
463 Finally, mitigation strategies related to feathering included implementing partial or full feathering
464 (pitching the turbine blades to 45 degrees or 90 degrees, respectively, to catch as little wind as possible
465 when below the cut-in speed for energy generation). In some cases, feathering was targeted towards high
466 activity periods, and it was primarily tested and implemented in the onshore wind industry. There was
467 substantial evidence of effectiveness from multiple studies for bats onshore, either on its own or in
468 combination with curtailment (American Wind Wildlife Institute, 2019; Arnett and May, 2016; Baerwald
469 et al., 2009). As with curtailment, there was a lack of testing and evidence for the effectiveness of
470 feathering as a mitigation measure for birds. There was also no evidence of effectiveness in the offshore
471 context.

472 3.3.3 Lighting to Reduce Attraction

473 There were 42 lighting-related mitigation approaches in the reviewed literature related to reducing
474 attraction. This included information from 82 source documents, with most sources discussing more than
475 one lighting alternative (E-TWG Bird and Bat Specialist Committee, 2020). Lighting-related mitigation
476 approaches that were suggested in the literature included limiting the use of lighting (n=18) as well as
477 altering various lighting characteristics such as color (n=6), visibility (n=2), flashing (n=3), and intensity
478 (n=2) to reduce attraction. Several visibility enhancement approaches (n=12) including those that are
479 lighting-related (e.g., using strobe light or ultraviolet light to make structures more visible such that
480 animals could avoid collisions) and non-lighting related (e.g., painting or texturizing turbine blades and
481 towers to increase visibility of wind farm structures) were also explored.

482
483 Studies of attraction of birds to various colors/schemes of light on offshore wind turbines have found that
484 it is best to restrict the light sources offshore as much as possible (Rebke et al., 2019). This included
485 approaches such as turning lights off when not in use, shielding lighting to only illuminate where needed
486 (particularly down-shielding), and using on-demand lighting of various types (e.g., timers, heat sensors,
487 motion sensors, ambient conditions; (E-TWG Bird and Bat Specialist Committee, 2020). Flood lighting
488 was broadly acknowledged to be hazardous to migrating birds, and so it is recommended that flood
489 lighting is avoided, particularly at night, in poor weather, and during migration (E-TWG Bird and Bat
490 Specialist Committee, 2020; Gauthreaux and Belser, 2006).

491
492 Limiting permanent lighting on offshore structures (as allowed by local safety regulations) was also a
493 commonly suggested mitigation approach. Aircraft Detection Lighting Systems (ADLS), for example,
494 were suggested numerous times in source documents to ensure that aviation obstruction lights on turbine
495 towers and nacelles were only lit when an aircraft was in the vicinity. All commercial-scale offshore wind
496 farms being planned in the United States have been required to implement ADLS (Bureau of Ocean
497 Energy Management (BOEM), 2024, 2023, 2022, 2021a). Using as few lights as possible - for example,
498 only lighting the outer turbines in an array, and using as few lights as necessary to light each structure -
499 was also suggested as a common-sense way to minimize the amount of artificial light to which wildlife
500 were exposed (Mockrin and Gravenmier, 2012). Triggering vessel navigation lighting around turbine
501 bases, rather than having such lighting operational during all periods of poor visibility (similarly to ADLS

502 for towers/nacelles), was also suggested as a way to minimize light pollution (Rebke et al., 2019), though
503 this approach has not been implemented in the source documents examined..

504

505 In addition to lighting reduction, the color spectrum is also a focus of lighting-related minimization
506 approaches. White light has been shown to attract birds more than other wavelengths of artificial light.
507 The color that leads to the least amount of attraction, however, is still debated in the literature. A study
508 using lampposts in the Netherlands, for example, found that 60-81% of nocturnally migrating birds were
509 disoriented and attracted by white lights, while fewer birds were influenced by red (54%), green (12-
510 27%), or blue (4-5%) lights (Poot et al., 2008). Poot et al. (Poot et al., 2008) hypothesized that shorter
511 wavelengths (blue and green light) were less disruptive to birds' magnetic compass used for migration
512 than red or white light, and several other studies in marine ecosystems in Europe seemed to support this
513 assertion (e.g., (Hill et al., 2014; van de Laar, 2007). However, a study in the terrestrial United States that
514 pointed lights into the night sky found that white, blue, and green lights all attracted birds, while red light
515 did not (Evans et al., 2007). Several studies suggested the use of red light including near Procellariid
516 breeding colonies (Syposz et al., 2021) and at offshore wind farms in the North Sea (Rebke et al., 2019)
517 Other studies at communications towers and terrestrial wind farms in the United States found that steady-
518 burning red lights attracted migrant birds, but flashing red lights did not (Kerlinger et al. 2010, Gehring et
519 al. 2009). While the available evidence clearly indicated that avoidance of the use of white lights helped
520 to mitigate impacts to birds, the best alternative lighting color is still unresolved, and may depend on
521 weather, taxon, and other variables. Much more limited data was available on bat responses to light. A
522 study on the coast of the Baltic Sea found that bat acoustic activity increased around red LED light as
523 opposed to white (Voigt et al., 2018). However, high levels of bat activity around the white beams
524 emitted by lighthouses has been noted for several decades (e.g., (Cryan and Brown, 2007; Stantec
525 Consulting Services Inc., 2016), and in a five-year study of bat mortalities at a terrestrial wind farm in the
526 United States, (Bennett and Hale, 2014) found that synchronized, flashing red aviation lights did not
527 increase bat mortality rates as compared to unlit turbines.

528

529 Other lighting characteristics have also been suggested as minimization approaches, including the use of
530 flashing lights and altering intensity. Flashing (but not strobing) light was found to reduce avian attraction
531 relative to continuously burning light (Cook et al., 2011; Evans et al., 2007; Gartman et al., 2016; Gehring
532 et al., 2009; Kerlinger et al., 2010) and has been widely recommended in both offshore and onshore
533 contexts, though most testing of this approach occurred onshore. One meta-analysis of avian collision
534 mortality from 30 terrestrial wind farms across North America found no significant difference between
535 mortality rates at turbines with red flashing aviation safety lighting and turbines without lighting at the
536 same wind farms (Kerlinger et al., 2010). The more the flash duration was reduced, and the longer the
537 time interval between flashes, the better for reducing wildlife interactions. Flash synchronicity (e.g.,
538 across all turbines in a wind farm) has also been suggested to be important (Bennett and Hale, 2014),
539 along with using the minimum intensity required for safety or other purposes (e.g., (Syposz et al., 2021)
540 and others). However, more research is needed on this topic.

541 *3.3.4 Visibility and Detectability Enhancement*

542 While in the instances described above, mitigation was focused on methods to reduce attraction to
543 artificial light, there were some instances where lighting was proposed to increase visibility of structures
544 to reduce potential for collision (n=12), many of which have been field tested for birds or bats (n=6) or

545 implemented with evidence of effectiveness for birds (n=4). The use of flashing or flickering lights were
546 tested in onshore environments for aero fauna. Day et al. (2005) determined that lights flashing 40
547 times/min at an island in Alaska significantly reduced nighttime flight velocity, resulted in a spatial
548 redistribution away from the island, and resulted in greater course change distances when visibility was
549 good or during tailwinds for Eider species (*Somateria* sp.). For bats, Gorresen et al. (2015) illuminated
550 trees with dim flickering UV light, which resulted in a reduction in Hawaiian hoary bat (*Lasiurus*
551 *cinereus semotus*) activity levels despite an increase in insects (prey), though there was a high level of
552 variation in results across turbines. However, a United States study of steady-burning dim ultraviolet
553 (UV) light used to illuminate terrestrial wind turbines found no difference in bat mortality rates between
554 illuminated and unilluminated turbines (Cryan et al., 2022). In contrast, the use of ultraviolet light to
555 illuminate electrical lines reduced avian collisions at those lines by 88% (Baasch et al., 2022). Previous
556 studies suggested that UV light does not attract and disorient birds as much as longer wavelengths (e.g.,
557 as cited in (Wiese et al., 2001), but additional data is needed on the utility of UV lights to illuminate
558 obstacles.

559
560 Other visual mitigation strategies explored for aero fauna involved painting or texturizing turbine blades
561 and towers to increase visibility of wind farm structures (Ahlén et al., 2007; American Wind Wildlife
562 Institute, 2019; Arnett and May, 2016; Hein and Straw, 2021). For bats, (Long et al., 2011) determined
563 that the common turbine colors ‘Pure White’ and ‘Light Grey’ were among those found to attract
564 significantly more insects than other colors tested, which was suggested to provide foraging opportunities
565 for bats and increase collision risk. Another study (Bennett and Hale, 2018) tested texturizing turbine
566 towers to reduce attractiveness to bats, and while lab and flight cage experiments indicated potential
567 effectiveness, initial field trials at wind turbines were inconclusive (Hein and Straw, 2021). Further
568 testing on these visual and sensory mitigation approaches is required to determine potential effectiveness
569 for bats, particularly in the offshore wind industry.

570
571 For birds, painting turbine blades was field tested and demonstrated evidence of effectiveness. An
572 experimental laboratory study on American Kestrels (*Falco sparverius*) indicated that painting a single
573 turbine blade black may have reduced motion smear and could be effective for reducing bird collision risk
574 (Hodos, 2003). This strategy was further tested at a wind farm in Norway, in which (May et al., 2020)
575 found that painting one blade black significantly reduced the annual bird mortality rate by over 70%
576 relative to neighboring unpainted turbines, with raptor mortalities experiencing the greatest reduction.
577 Other research supported the importance of increasing contrast of turbine blades to improve bird’s
578 rotation speed detection (Blary et al. 2023). Painting turbine foundations has also shown effectiveness for
579 mitigating bird collisions (Gartman et al., 2016). A field study investigating the impact of tower base
580 painting on Willow Ptarmigan (*Lagopus lagopus*) mortality rates found a 48.2% reduction in the annual
581 number of recorded carcasses per search at painted relative to unpainted turbines (Stokke et al. 2020). As
582 some birds have the ability to see in the UV spectrum, using UV-reflective paint to increase turbine blade
583 and tower visibility was also suggested. One paper recommended, but did not test, the use of highly
584 reflective achromatic patterns (black and white) that extended across the full spectrum of incident light
585 (Martin and Banks 2023). Although UV paint is an established method for reducing bird strikes against
586 windows (Marques et al. 2014, Defingou et al. 2019), its effectiveness in onshore and offshore wind
587 development has not been proven (Young et al. 2003).

588 3.3.5 Vessel Operations and Water Quality

589 In addition to approaches related to turbine operations (e.g., curtailment, feathering), and lighting and
590 visibility of turbines and other structures, there are additional operational considerations (n=12
591 approaches) focused on vessels and safety measures to ensure water quality. These measures were
592 primarily relevant to birds; for a majority (n=10), the implementation status was unknown or not
593 implemented, while two approaches in our literature review were field tested. Vessels in the marine
594 environment have the potential to disturb wildlife, including marine birds (Schwemmer et al., 2011). Six
595 mitigation approaches related to vessel operation parameters were reviewed, which include those aimed at
596 reducing bird disturbance by minimizing vessel speed and density, implementing setbacks from high-use
597 areas (i.e., breeding or foraging grounds), and using caution in low visibility conditions. Limiting vessel
598 speed and volume and adopting set-back distances were field tested as mitigation strategies in maritime
599 industries. For example, (Bellefleur et al., 2009) determined that Marbled Murrelets required a minimum
600 buffer distance of 29 m between a boat and birds on the water to reduce disturbance such that 75% of the
601 studied population would be minimally affected. Additionally, (Ronconi and St. Clair, 2002)) found that
602 using a vessel setback distance of 600 m from shore with a speed limit of 25 km/h reduced Black
603 Guillemot (*Cepphus grylle*) flushing behavior to 10% most of the time. At onshore wind farms, there has
604 also been some suggestion to manage vehicle volumes and remain particularly alert when visibility is
605 poor to reduce aero fauna collisions and disturbance (Mockrin and Gravenmier, 2012; U.S. Fish and
606 Wildlife Service (USFWS), 2012), but this has not yet been tested or implemented. Additionally, no
607 studies examined the effectiveness of implementing vessel operation parameters for offshore wind
608 development specifically.

609
610 Water quality is tied to the overall health of the ecosystem. There were five mitigation strategies tied to
611 water quality from three sources, but none of these mitigation strategies were tested for effectiveness.
612 Potential strategies include implementing methods to reduce impacts to water quality (i.e., best
613 management practices), following applicable regulations, proper handling of toxic substances following
614 appropriate federal and state measures, and reducing the likelihood of invasive species introduction (U.S.
615 Fish and Wildlife Service (USFWS), 2012).

616 3.3.6 Deterrents and Attraction Reduction

617 Thirty-seven different mitigation strategies were proposed from 69 literature sources related to deterring
618 aero fauna from approaching activities and/or structures or reducing the attractiveness of such activities
619 and/or structures. Of these, 17 were untested (or their testing status could not be determined), another 10
620 were field tested, and 9 were implemented with evidence of effectiveness. While some industries
621 distinguish between deterrents and hazing devices, for the purposes of this review, we use ‘deterrents’ as
622 an encompassing term to include both. There are various types of deterrents aimed at scaring or otherwise
623 repulsing individuals to counteract attraction and in turn reduce collision risk. These act on various senses
624 of individuals and include acoustic (n=9), visual (n=6), physical (n=4), olfactory and chemical (n=2)
625 deterrents, or some combination of these (n=5). In addition, approaches to reduce attraction have been
626 suggested that relate to altering prey resources (n=7) and habitat (n=4).

627
628 Various types of acoustic deterrents have been suggested and tested. Ultrasonic deterrents (which jam the
629 ability of bats to receive and interpret echolocation calls) have shown some evidence of effectiveness for

630 reducing bat mortality at onshore wind facilities (Hein et al., 2021; O'Neil, 2020). More specifically, one
631 study found that broadband ultrasound in the 20-100 kHz range decreased bat mortality by up to 64% per
632 treatment turbine relative to control turbines, but that effectiveness is limited by the distance and area that
633 ultrasound can be broadcast (Arnett et al., 2013)). Similarly, research conducted by (Gilmour et al., 2020)
634 determined that ultrasonic auditory deterrents lowered bat activity by 80% when deployed alone and in
635 combination with radar at riparian sites in Wales. In a related study, also in Wales (Gilmour et al., 2021),
636 auditory deterrents significantly reduced passes of Soprano Pipistrelle (*Pipistrellus pygmaeus*; 27%),
637 Mouse-eared bats (*Myotis* spp.; 29%), and Noctule (*Nyctalus* spp.) and House bats (*Eptesicus* spp.; 68%).
638 However, research performed in an agricultural setting in the United States did not detect a clear
639 reduction in bat mortalities from the use of ultrasonic acoustic deterrents for any individual species,
640 including Hoary Bats (*Lasiurus cinereus*), Silver-haired Bats (*Lasionycteris noctivagans*), and Big Brown
641 Bats (*Eptesicus fuscus*; (Schirmacher, 2020)). The mortality rate of Eastern Red Bat (*Lasiurus borealis*)
642 was also estimated to be 1.3-4.2 times greater when turbines were operating normally and deterrents were
643 on, as compared to when deterrents were off. Eastern red bats are the most common species detected
644 offshore in the northwest Atlantic, and ultrasonic deterrents have not yet been tested in the offshore
645 context for them or any other species.

646
647 Several studies from Europe in the literature review tested the use of other auditory devices, such as
648 electromagnetic signals, to deter bats from approaching wind turbines, with some evidence that the
649 approach reduces bat activity levels (Nicholls and Racey, 2009, 2007). For example, Nicholls and Racey
650 (2009) found that bat activity levels and foraging effort in Scotland were both reduced during
651 experimental trials with exposure to a unidirectional radar antenna. Monitoring efforts around traffic
652 control, military, and weather radar showed reduced bat activity in areas with electromagnetic exposure of
653 strength of >2 volts/meter for multiple *Pipistrellus* and *Myotis* bat species (Nicholls and Racey, 2007),
654 suggesting that this may be a potential approach for further study in the offshore context.

655
656 In the case of birds, bioacoustic deterrents, which primarily involve the playback of recorded alarm or
657 distress calls, have been effectively used in agricultural and riparian settings (Enos et al., 2021), though
658 effectiveness appears to vary by species, and habituation is common. However, according to a literature
659 review on bioacoustics (Bomford and O'Brien, 1990), more than 80% of visiting Black-crowned Night
660 Herons (*Nycticorax nycticorax*) were frightened off trout ponds when recorded distress calls of juvenile
661 and adult Night Herons were broadcast, and habituation did not occur after 6 months. In another study,
662 cornfields receiving taped Carrion Crow (*Corvus corone*) distress calls had significantly less damage than
663 the control or fields with suspended dead crow bodies (Bomford and O'Brien, 1990). Several lab
664 experiments testing bioacoustic effects on starlings determined that bioacoustic devices cause an aversive
665 response when compared to other noises (Bomford and O'Brien, 1990), though habituation was reported.
666 In contrast, recorded alarm calls were ineffective at deterring Canada Geese (*Branta canadensis*) in
667 recreational areas (Aguilera et al., 1991).

668
669 Other studies have examined using loud noises (e.g., pyrotechnics, bangers, screamers, exploders) to deter
670 birds from an area with varying effectiveness. Propane exploders reduced Red-winged Blackbird
671 (*Agelaius phoeniceus*) damage to cornfields by 77%, with no evidence of habituation after 4 weeks
672 (Conover, 1984). Screamer shells reduced the number of geese in urban parks for up to 15 days
673 ((Aguilera et al., 1991). In contrast, response to pyrotechnics at a landfill site was species-specific, with a

674 reduction in Herring Gull (*Larus argentatus*) numbers, but little response by crows and starlings (Curtis et
675 al., 1993). In general, bird habituation to loud noises may limit deterrent effectiveness over longer time
676 periods. Presenting sounds at random intervals, using a range of different sounds or moving them
677 frequently, and supporting sounds with additional auditory or visual methods can help to increase long-
678 term effectiveness (Bomford and O'Brien, 1990; Gilsdorf et al., 2002).

679
680 In the offshore environment, aquatic hazing devices were ineffective at deterring waterbirds species, such
681 as Greater and Lesser Scaup (*Aythya affinis* and *A. marila*) and Surf Scoter (*Melanitta perspicillata*), from
682 oil spills in the San Francisco Bay (Whisson and Takekawa, 2000). Other research suggested that using
683 acoustic whistle cues (2-4 kHz) may help birds hear turbine blades and avoid collision, but this strategy
684 was not tested (Dooling, 2002). Lastly, ultrasonic acoustic deterrents were implemented as a strategy to
685 deter birds from perching at offshore platforms (Christensen-Dalsgaard et al., 2019), but there was
686 minimal empirical proof of their effectiveness (Defingou et al., 2019).

687
688 Visual deterrents are a common strategy used in agriculture and other terrestrial industries (e.g., airports,
689 landfills), and therefore many have been field tested and implemented. Visual deterrents have included
690 lasers, effigies, flagging, flickering light, unmanned aerial vehicles (UAVs), automated detection systems,
691 and paint. Lasers have deterred certain bird species under low light conditions but were not consistently
692 effective across studies and did not disperse birds long distances. For example, in field trials, (Glahn et
693 al., 2000) directed lasers at roost trees at or near sunset, which dispersed Double-crested Cormorants
694 (*Phalacrocorax auritus*) and reduced roost populations by at least 90% after 1 to 3 evenings of
695 harassment. However, there have been variable results for other species, including Canada Geese, with
696 some instances of effectiveness at dispersing birds (Blackwell et al., 2002) and other efforts being
697 ineffective (Sherman and Barras, 2004). In the maritime environment, automated laser systems were
698 successfully used on installations to deter gulls from large open areas such as helicopter landing
699 platforms, and seemed to work best in low light conditions (Christensen-Dalsgaard et al., 2019). For
700 offshore wind development, lasers were identified as a potential mitigation method to minimize avian
701 collision risk with turbines in the UK (Cook et al., 2011), but this approach was not tested.

702
703 Effigies, including scarecrows, hawk-kites, and eye-spot balloons, have also been used as a visual
704 stimulus to deter birds (Arnett and May, 2016; Enos et al., 2021; Gilsdorf et al., 2002). (Conover, 1984)
705 reported that hawk-kites, plastic kites imprinted with a picture of a flying hawk and suspended in the air,
706 reduced Red-winged Blackbird damage to cornfields by 83%, with no habituation observed during the 4-
707 week study period. (Stickleby and King, 1993) found that the use of a "Scary Man" pop-up inflatable
708 effigy device proved effective in reducing Double-crested Cormorant numbers on catfish ponds for 10-19
709 days. Eye-spot balloons (floating balloons mimicking a generic aerial predator) have also reduced
710 passerine abundance and/or resulting crop damage in several studies conducted at vineyards and
711 agricultural fields, but more research is required to determine effectiveness (Enos et al., 2021).

712
713 Reflective materials have been tested as a deterrence method for birds in terrestrial contexts, with variable
714 results. Mylar reflective tape was ineffective at deterring Herring Gulls from nesting colonies, but did
715 reduce the use of loafing sites in some cases (Belant and Ickes, 1997) was ineffective in reducing bird
716 damage to agricultural fields, possibly due to habituation (Gilsdorf et al., 2002). However, white plastic

717 flags successfully repelled Snow Geese (*Chen caerulescens*) from rye and winter wheat fields during a
718 two-month study period (Gilsdorf et al., 2002).

719
720 UAVs, which include drones or remote-controlled helicopters, have been found to significantly reduce
721 crop damage and bird abundances in several experiments on passerines in vineyards (Enos et al., 2021).
722 For bats, a 3-week study indicated that a UAV deterrent significantly reduced bat activity in the relevant
723 airspace of up to 700m above the UAV's flight altitude (Werber et al., 2023). There have not been any
724 studies on the effectiveness of UAVs as deterrents in relation to onshore or offshore wind development
725 and there may also be considerations to avoid disturbance from these methods.

726
727 Structural deterrents, such as wires, spikes, and netting, were tested as a method to mitigate avian
728 perching or roosting behavior on onshore and offshore wind turbines and other infrastructure, but their
729 effectiveness was unknown (Curry and Kerlinger 1998, Clarke 2004, Cook et al. 2011). Available
730 research suggested that physical deterrents may work best when installed on horizontal structures (i.e.,
731 staircases, railings) and be more effective for certain species, including raptors and gulls (Christensen-
732 Dalsgaard et al. 2019). Decoy towers without functioning rotor blades, positioned at the end of turbine
733 rows, were also tested as a potential method to deter raptor species from perching at an onshore wind
734 facility (Curry and Kerlinger 1998). However, their effectiveness was not determined. While these
735 alternative perches may have helped keep birds off of active turbines, it may have also increased bird
736 attraction to the general area of the turbines and heightened collision risk (Smallwood and Karas 2009,
737 Cook et al. 2011).

738
739 Several studies tested chemical deterrents (taste, olfactory, and tactile repellents) to deter birds from
740 landfills, crop fields, and other onshore areas (Gilsdorf et al. 2003, Cook et al. 2011). Curtis et al. (1993)
741 evaluated the impacts of methyl anthranilate and found that the treatment did not effectively reduce avian
742 numbers at a landfill site. Comparatively, Conover (1984) determined that using 4-amino-pyridine
743 (Avitrol FC-99) on corn fields increased the number of Red-winged Black Birds, but decreased the
744 number of sparrows following treatment, suggesting chemical responses may be species-specific. In the
745 offshore industry, the use of 'Bird Free Gel', a sticky gel emitting UV light and a deterring smell and
746 taste, was implemented on an offshore oil installation in the North Sea and used successfully in other
747 areas to deter breeding Kittiwakes (*Rissa tridactyla*), gulls, and roosting pigeons (Christensen-Dalsgaard
748 et al. 2019). However, other research noted that chemical repellents had little specificity to reducing
749 offshore and onshore wind turbine collision risk, and their permanence over time may depend on terrain
750 and weather conditions (Marques et al. 2014, May et al. 2015, Arnett and May 2016). In general, there
751 was little evidence of effectiveness for any of these deterrents in the offshore context, and the level of
752 effectiveness varied based on species, environmental conditions, and other factors.

753
754 In addition to the use of a particular type of deterrent, several studies assessed the integration of various
755 deterrent techniques. For example, Montoney and Boggs (1995) used a combination of pyrotechnics,
756 amplified distress tapes, and live ammunition to harass Laughing Gulls (*Leucophaeus atricilla*) away
757 from an airport, resulting in an 86% reduction in aircraft strikes and a 100% decrease in runway closures
758 due to bird activity. For Canada Geese, Mott and Timbrook (1988) determined that a combination of
759 alarm calls and racket bombs produced a 96% reduction in geese observations on lawns, though continual
760 harassment would likely be necessary as reinvasion of geese was observed one week after treatment

761 stopped. An integrated approach using alarm calls, pyrotechnics, and methyl anthranilate was shown to be
762 effective at reducing waterfowl landings at ponds with no apparent habituation (Stevens et al. 2000).
763 Combining different deterrent strategies has been tested for mitigating bird collisions with terrestrial wind
764 turbines as well. A review paper (Garcia-Rosa and Tande 2023) highlighted results of a 9-month study
765 that tested a deterrence module operation for raptors at an operational wind farm, which involved species-
766 level identification, unmanned aerial vehicles, and seven detection and deterrent systems. It was estimated
767 that the installed systems potentially reduced Golden Eagle collision risk by 33–53%. Yet, a pilot study at
768 a wind turbine facility in Nova Scotia Canada determined that using predator owl deterrent models and
769 bioacoustic alarms did not significantly deter birds from wind turbines (Dorey et al. 2019). As such, the
770 effectiveness of integrated deterrent approaches for birds at onshore wind farms remains unclear, and in
771 our review of the literature, no similar approaches were tested at offshore wind farms.

772
773 Lastly, efforts to reduce bird attraction to certain structures or areas have used habitat enhancement in
774 other locations. For instance, it has been suggested that supplementary or diversionary feeding could be
775 effective in reorienting foraging harriers away from profitable prey areas (Simmons et al. 2020).
776 However, while a study of Hen Harriers (*Circus cyaneus*) at an onshore wind facility showed an increase
777 of 32–42% in flight activity in a created habitat enhancement area next to a terrestrial wind facility, there
778 was no difference in activity within the wind facility (Gartman et al., 2016). In contrast, the felling of a
779 plantation forest increased Golden Eagle use of the area, with a shift in their range away from the wind
780 farm (Gartman et al., 2016). Reducing the attractiveness of the wind farm itself has also been tested with
781 some effectiveness. For example, superficially tilling the soil (3–8 cm deep) at the base of wind turbines
782 in Spain made the area less attractive for Lesser Kestrels (*Falco naumanni*; Garcia-Rosa and Tande
783 2023). Similarly, rock piles and other areas attractive to rodent prey were correlated with raptor mortality
784 at turbines at Altamont Pass Wind Resource Area in California, and it was suggested that making the
785 areas around turbines less attractive to raptor prey could help to reduce mortalities (Thelander et al.,
786 2003). However, more research is needed in this area, particularly in terms of relevant approaches for the
787 offshore wind industry.

788 3.4 Compensation

789 Compensation for environmental effects that cannot be avoided or minimized may occur onsite or offsite
790 and may include both in-kind or out-of-kind compensation (e.g., providing a resource of similar or
791 different structural or functional type to the impacted resource). Federal regulations, such as the National
792 Environmental Policy Act (NEPA) in the United States and the Birds and Habitats Directives in the
793 European Union (Van Hoorick, 2014), provide regulatory frameworks for compensatory mitigation of
794 impacts to species and their habitats, but offsets can also be pursued outside of federal permitting
795 processes. Compensation has generally been implemented via three primary mechanisms: 1) direct
796 implementation by an individual project, 2) mitigation banking, or 3) in-lieu fee programs. Direct
797 implementation (often via permittee-responsible compensation) includes the implementation of shovel-
798 ready projects such as habitat restoration or creation. A mitigation bank is a site or suite of sites that
799 provides ecological functions and services expressed as credits that are conserved and managed in
800 perpetuity for particular species and are used expressly to offset impacts occurring elsewhere to the same
801 species. In-lieu fee programs, which are similar to mitigation banks, provide ecological functions and
802 services expressed as credits that are conserved and managed for particular species or habitats and used to
803 offset impacts occurring elsewhere to the same species or habitats (USFWS 2023).

804

805 We reviewed sixteen papers related to compensation as a means of mitigating unavoidable impacts. The
806 majority (75%) of these studies looked at compensation methods involving in-kind habitat conservation,
807 restoration, restriction of access (either from fishing or recreation), invasive species management,
808 reduction of bycatch, or other methods. The remaining studies (25%) reviewed financial means of
809 compensation either through development of funding pools, permittee-required compensation strategies,
810 financial support of rehabilitation programs, or in-lieu fee programs.

811

812 The majority of compensation strategies did not provide examples of implementation or assess the
813 effectiveness of compensation programs (n=13). There was little documented evidence within the
814 reviewed literature showing effectiveness of financial compensation at off-setting unavoidable impacts
815 related to offshore wind development. In part, this may be due to the relatively nascent status of the
816 industry in many parts of the world. Compensation has been used within the United States regulatory
817 context for other types of projects, primarily terrestrial, including Clean Water Act compliance (Arnett
818 and May 2016), and often includes habitat protection and restoration, which may be difficult to
819 measurably implement in the marine environment. Once the offshore wind industry progresses and such
820 projects have been implemented, a thorough review of the results of these strategies would provide better
821 understanding of the role compensation can play in mitigating impacts on aerofauna.

822

823 One strategy that has been extensively tested (though not specifically in relation to OSW development)
824 focused on the removal of mammalian predators as a means of increasing productivity of seabird breeding
825 colonies. Removal of invasive predators from breeding colonies has generally been found to be an
826 effective strategy (e.g., (Nogales et al., 2013; Pascal et al., 2008). Results from our reviewed studies were
827 somewhat mixed; Luxmoore et al. (2019) found that Guillemot/Murre (*Uria* spp.) populations did not
828 respond to the removal of rats, while Booker et al. (2019) attributed increases in Razorbill (*Alca torda*)
829 and Atlantic Puffin (*Fratercula arctica*) breeding numbers to the removal of rats (Booker et al. 2019).

830 **4. Discussion**

831 There were few mitigation measures available for birds and bats at offshore wind farms that were
832 adequately tested in the reviewed literature. Commonly implemented measures often had not been
833 demonstrated to be effective (e.g., raising the lower end of the rotor swept zone to reduce marine bird
834 collision risk) or had been shown to be effective in other scenarios such as with terrestrial wind energy
835 development, but not proven offshore (e.g., curtailment of turbine operations based on environmental
836 conditions to reduce bat collision risk). Available data suggested that in order for mitigation to be
837 effective, it likely needs to be species- and location-specific: a measure that works at one site for a given
838 species and life history stage may not be as effective in other contexts. Thus, even the measures that were
839 successfully used in the terrestrial context (or with other marine industries such as offshore oil and gas)
840 should be tested in relation to OSW development and particular species of interest, to ensure that the
841 measures can effectively reduce risk. Since it may be difficult to identify mitigations that are broadly
842 generalizable, an array of mitigation measures should be further explored to provide options for specific
843 scenarios.

844

845 Despite a lack of adequate testing offshore, there are a few key mitigation measures that we suggest could
 846 be implemented across offshore wind facilities without the need for further study and testing, due to a
 847 strong base of evidence from other industries and/or because they are common sense measures that could
 848 be easily implemented. Other approaches may be promising, with limited evidence or in some cases
 849 conflicting evidence, and thus require additional research before consideration for offshore
 850 implementation. Still other methods either lack evidence (and thus also require research to assess their
 851 effectiveness) or are likely ineffective or inappropriate to implement in the offshore context. We have
 852 identified existing mitigation approaches that fall into each of these categories (Table 1) to help guide
 853 future mitigation implementation and research, recognizing that this is a very general assessment, and a
 854 one-size-fit-all approach does not exist for mitigation.

855 4.1 Limitations of the Literature Review

856 Due to the search terms used in our review, and how these terms are generally used in the literature, our
 857 literature review was primarily focused on minimization measures, rather than on avoidance or
 858 compensation. We suspect that this review represents a gross underestimate of the true implementation
 859 frequency for these types of mitigation measures. Despite the prevalence of the mitigation hierarchy in the
 860 **Table 1.** Categories of mitigation approaches for aero fauna at offshore wind farms and evaluation of available
 861 evidence and research needs. Mitigation approaches were placed into four categories: **green** = evidence of
 862 effectiveness OR a common sense approach to avoid conflict, but uncertain evidence due to logistical difficulty of
 863 study, **yellow** = some evidence of effectiveness for specific approaches or circumstances, but more research needed,
 864 **orange** = conflicting evidence, more research needed, **red** = no evidence it is effective, or there is evidence of
 865 ineffectiveness, **gray** = not relevant/appropriate.
 866

| Hierarchy Category | Mitigation Approach | Birds | Bats |
|---------------------------------|--|--------|--------|
| <i>Avoidance</i> | Siting considerations | Green | Green |
| | Seasonal approaches | Green | Green |
| | Limiting activities | Green | Green |
| | Engagement and communication | Green | Green |
| <i>Minimization</i> | Turbine layout | Yellow | Yellow |
| | Turbine size | Orange | Red |
| | Curtailement | Orange | Yellow |
| | Feathering | Red | Yellow |
| | Temporary shutdown | Yellow | Red |
| | Limiting use of lighting | Green | Green |
| | Altering light color spectrum | Green | Orange |
| | Altering other lighting characteristics | Green | Red |
| | Visibility/detectability enhancement | Yellow | Yellow |
| | Use of acoustic deterrents | Orange | Yellow |
| | Use of visual deterrents | Orange | Red |
| | Use of physical deterrents | Yellow | Red |
| | Use of chemical and olfactory deterrents | Orange | Red |
| | Use of combined deterrents | Orange | Red |
| | Other attraction reduction | Red | Red |
| | Altering vessel speed and distance | Yellow | Gray |
| Water quality | Green | Gray | |
| <i>Compensation and Offsets</i> | Implementing compensation | Yellow | Yellow |

867
868 scientific literature, the term “mitigation” is also sometimes used synonymously with “minimization,” and
869 that was evident in the types of studies that were located during our review. Thus, the scope of this review
870 largely excluded marine spatial planning exercises, stakeholder engagement activities, and permitting
871 documents developed by governmental agencies, which was where most siting-focused and
872 compensation-focused mitigation tended to be discussed. Compensation for OSW development
873 specifically is also a relatively recent topic of discussion in the scientific community (Croll et al., 2022).
874 Finally, some of these activities (e.g., stakeholder engagement by government entities) also may not have
875 been particularly well represented in the scientific literature, despite our best efforts to include gray
876 literature.

877
878 In addition to the information provided in this paper, other data on the effectiveness of various mitigation
879 measures can be found via the website Conservation Evidence (<https://www.conservationevidence.com/>).
880 While not offshore wind-specific, this international project takes a science-based approach to assessing a
881 wide range of biodiversity conservation measures and includes links to the scientific studies where
882 mitigation effectiveness was examined. Several other efforts have focused specifically on assessing the
883 effectiveness of mitigation measures for OSW development, though many of these have not focused on
884 aeroфаuna (for example, see (Verfuss et al., 2016). One recent assessment did include birds to some
885 degree (Crown Estate Scotland, 2024). This report aimed to inform OSW development in Scotland by 1)
886 identifying the “most appropriate mitigation measures” for OSW developments in this region, and 2)
887 examining approaches for Nature Inclusive Design at Scottish projects. However, it should be noted that
888 this assessment was unable to find data on the effectiveness of mitigation measures in most cases, and
889 thus chose to define mitigation success as situations in which “no deleterious effects were recorded” and
890 therefore evidence of effectiveness as defined in this paper is lacking. This speaks to the difficulty in
891 assessing effectiveness of mitigation approaches and general lack of data, however we did our best to
892 assess the available information and scientific evidence for each approach in order to provide additional
893 utility in identifying mitigation measures that should be implemented versus those that require additional
894 study.

895 **4.2 Approaches with Evidence of Effectiveness**

896 *4.2.1 Avoidance*

897 Despite a lack of direct research in our literature review on the effectiveness of siting and other avoidance
898 strategies, we strongly recommend implementation of these mitigation approaches. Avoidance-focused
899 mitigation strategies are likely harder to test than some other type of mitigation such as on-site
900 minimization measures, leading to relatively little evidence of effectiveness in the literature, despite the
901 common-sense value in avoiding wildlife interactions as a way to mitigate effects. Nevertheless, there is a
902 strong scientific consensus, as far back as 2003, that siting and other avoidance measures remain the best
903 available approach to reduce OSW development impacts, particularly for birds (Langston and Pullan,
904 2003).

905
906 In addition to siting and limiting the timing of certain activities, we also recommend implementing
907 mitigation strategies related to water quality. While there were relatively few mitigation approaches
908 related to water quality identified in the database, these are generally common-sense approaches that

909 should be implemented to avoid potential negative impacts. This includes practices like cleaning
910 vehicles/equipment to reduce the likelihood of invasive species introduction and following safety
911 measures to avoid the introduction of toxic substances (Mockrin and Gravenmier, 2012).

912 It is generally difficult to test the effectiveness of avoidance measures, because there are typically no
913 “impact” scenarios to compare to “controls” and the mitigation generally happens prior to the project
914 even being built. As OSW farms continue to be built around the world, it will be important to gather data
915 on avoidance-based mitigation measures via natural experiments (e.g., as projects are sited in locations
916 with varying value to wildlife, and resulting impacts can be compared across projects). Continued
917 development will also help to inform our understanding of what makes a “good” or “poor” location from
918 the standpoint of avoiding aerofauna impacts; for terrestrial wind energy, for example, our understanding
919 has shifted over time regarding what types of sites (based on landscape features such as topography and
920 land cover, as well as measurements of wildlife activity) will lead to more or fewer aerofauna impacts. As
921 the scientific community has discovered that pre-construction acoustic activity levels of bats, for example,
922 or passage rates of migrants measured via marine radar, do not necessarily predict collision risk at
923 terrestrial wind farms there is a need for the scientific and management communities to explore
924 alternative ways of understanding what factors will most accurately predict post-construction effects,
925 including mortality rates. This type of ongoing assessment to inform the adaptive management of future
926 projects is an essential component of mitigation efforts.

927 *4.2.2 Minimizing Attraction*

928 While uncertainty remains regarding the best colors of lighting to reduce attraction of aerofauna, there is
929 strong evidence for the benefits of an overall reduction in lighting during all phases of OSW
930 development, including minimizing the use of white lights where possible. Artificial lighting can have
931 serious impacts on wildlife populations, species, and ecosystems (Jägerbrand and Bouroussis, 2021;
932 Walsh et al., 2024). All OSW facilities should enact mitigation measures to reduce artificial light at night
933 (ALAN) pollution and avoid attracting and disorienting birds, particularly nocturnal migrants. This
934 includes minimizing the use of steady-burning and white lights whenever possible, down-shielding of
935 lights (but also not shining light directly on the water’s surface), and other measures that have been
936 detailed in several guidance documents for OSW development (Bureau of Ocean Energy Management
937 (BOEM), 2021b; E-TWG Bird and Bat Specialist Committee, 2020) as well as other industries (e.g.,
938 Environment and Climate Change Canada, 2016). While lights are required on turbines and other
939 offshore infrastructure for aviation and marine navigation safety purposes, and regulations vary by
940 country, it is recommended that the use of aircraft detection lighting systems (ADLS) and similar
941 mechanisms to reduce ALAN be implemented wherever possible, including adjustments of current
942 regulations where needed to allow for such measures. Additionally, temporary lighting may be a
943 particular risk; reduction in use of flood-lighting during construction and maintenance activities,
944 especially during spring and fall migration, is essential to minimize risks to migrating birds (E-TWG Bird
945 and Bat Specialist Committee, 2020).

946
947 Recent guidance for lighting of OSW developments in the United States recommends the use of many of
948 the above mitigation approaches to reduce the effects of artificial light on birds, bats, and other wildlife
949 (Bureau of Ocean Energy Management (BOEM), 2021b; E-TWG Bird and Bat Specialist Committee,
950 2020). This includes the use of red aviation safety lighting at the lowest practicable flash rate, avoidance

951 of flood lighting where possible, and the use of timers and other sensors to enforce lighting shutoffs. All
952 projects in the United States that have received final permitting approval to date for construction and
953 operations plans have been required to use ADLS to further reduce lighting of turbine structures (see, for
954 example, (Bureau of Ocean Energy Management (BOEM), 2024, 2023, 2022, 2021a), which has
955 environmental as well as social benefits (e.g., by reducing the nighttime visibility of turbines from shore).
956 New OSW developments in other regions would do well to take a similarly proactive approach to lighting
957 offshore structures.

958
959 The use of alternative lighting options to reduce impacts to birds is perhaps the best-understood and most
960 thoroughly tested mitigation category reviewed. Use of non-white, flashing, less intense, shielded lights -
961 with those lights operating as infrequently as possible - is the best available approach to reduce avian
962 attraction, disorientation, and mortality around offshore wind structures. Nevertheless, additional research
963 would be beneficial to better understand which wavelengths of (non-white) light can best reduce risk for
964 different weather conditions and taxa, since available studies are contradictory on this point. A better
965 understanding of bat response to light would also be useful, as uncertainty remains regarding the
966 mechanisms of attraction for this taxon. There is some evidence that bats actively approach turbines, and
967 this may be dictated by multiple factors including potential foraging opportunities (Guest et al., 2022). It
968 is possible that bat attraction may be dictated less by their own reaction to light than by the attraction of
969 their insect prey or other factors (Guest et al., 2022). Future aerofauna studies should be focused
970 specifically on the offshore environment to best inform mitigation for OSW development. Additional
971 study of aerofauna responses to UV light would help inform its use as a possible mitigation measure. Our
972 relatively poor understanding of animals' vision likely hinders our ability to make informed decisions
973 about mitigation in this area.
974

975 **4.3 Approaches that Require Additional Research**

976 Many potential mitigation strategies that have been identified in this review require additional research
977 and testing before consideration for widespread implementation in the offshore context. This includes
978 approaches that may have 1) some existing evidence of effectiveness for a specific approach or
979 circumstance, 2) conflicting evidence, or 3) have yet to be tested. All approaches relating to the
980 configuration of structures, deterrents, and altering vessel parameters fall into these categories.
981

982 *4.3.1 Minimization Through Project Design*

983 The layout of offshore wind farms and the specifications of turbines (including size and air gap) may
984 influence both how animals perceive the wind farm as a whole and how they move in and around the
985 wind farm. This can have direct consequences for attraction, avoidance, and risk of collisions for
986 aerofauna. While there is some evidence of the importance of turbine size and air gap in the terrestrial
987 wind context, a great deal of additional research is needed to understand the degree to which these factors
988 affect aerofauna in the offshore context. Understanding the effects of structure configuration on indirect
989 effects such as behavioral disturbance, displacement, and habitat fragmentation can be challenging, as this
990 requires regional-scale studies across multiple wind farms to allow for comparison. This type of multi-
991 facility study requires consistent data collection methods across projects along with public availability of

992 data from site-specific studies. Detecting changes in behavior may also require advanced technology or
993 human observers, both of which can be costly to deploy at the large scale required. Understanding the
994 effects of structure configuration on collisions offshore, which are both rare and difficult to detect,
995 likewise requires a high level of effort to achieve adequate statistical power. Mitigation through project
996 design offers promise to mitigate potential effects to aerofauna from OSW development, but this promise
997 remains largely unproven and will require large-scale, long-term research efforts. We suggest that, if such
998 approaches are implemented, they are accompanied by a rigorous regional-scale monitoring plan to assess
999 their effectiveness.

1000 *4.3.2 Minimizing Attraction and Enhancing Visibility*

1001 Based on findings from other industries and contexts, there are many types of deterrents that offer some
1002 level of promise. However, in many cases the studies of effectiveness focused on one or a few species
1003 over a short period of time, and often showed variable results. While deterrents have been used in the
1004 OSW industry in Europe to reduce perching for taxa such as gulls and cormorants (Gartman et al., 2016),
1005 these facilities were located closer to shore than those being built in North America. Consequently, there
1006 is uncertainty in the level of attraction (for roosting, perching, and nesting) likely to be observed in the
1007 North American context, and thus the need for these types of deterrents. As discussed above, the use of
1008 offshore wind turbines by bats remains largely unknown in the region as well. Thus, for this type of
1009 mitigation, we recommend not only additional research on effectiveness but also to only consider
1010 implementation once there is evidence that attraction is occurring at large enough scales to warrant
1011 investment in deterrents. While in general it is thought that reducing perching and roosting opportunities
1012 can help reduce attractiveness of wind farms to birds, which in turn may help reduce collision risk
1013 (Gartman et al., 2016), we lack direct evidence linking perching deterrents to collision risk reduction. Key
1014 studies to fill knowledge gaps include those specific to species that occur offshore and long-term studies
1015 to understand the potential for habituation.

1016
1017 There are other approaches to minimize attraction (and collision risk) that offer promise but also require
1018 additional testing and validation. The use of paint for increasing visibility of turbines and other structures
1019 shows promising evidence in other contexts but has to date been tested in a small number of contexts and
1020 with few species, all in the onshore context. Differences in vision among bird (and bat) taxa suggest that
1021 vision-related measures to enhance turbine visibility must be carefully considered and tested.

1022 *4.3.3 Minimizing Collisions*

1023 One of the largest data gaps identified in this review (as well as others; (Hein et al., 2021) is the degree to
1024 which bat mortality at offshore wind farms needs to be mitigated. Because of the high mortality rates of
1025 migratory tree bats at onshore wind facilities (Arnett et al., 2008), there have been serious concerns about
1026 the long-term impacts to these populations (Frick et al., 2017), and mitigation measures for terrestrial
1027 wind farms have received a high level of attention. In particular, curtailment of onshore turbines at low
1028 wind speeds has been shown to significantly reduce mortality rates, both when examining this approach at
1029 an individual project scale (Arnett et al., 2008) and when looking across projects (Adams et al., 2021;
1030 Whitby et al., 2021). While there are many variations in how curtailment is implemented (e.g., blanket
1031 curtailment as well as curtailment based on environmental conditions, temporal variation, smart systems
1032 that detect bats, etc.), the aim is the same: to stop turbines from spinning when bats are in the area.

1033 Indeed, feathering and temporary shutdown, while separate mitigation approaches, are also effective as
1034 they also result in the stoppage or minimization of turbines spinning, resulting in reduced collisions for
1035 bats. Much of the reason for variations in this type of mitigation approach, and the focus on conditional
1036 implementation, has more to do with the loss of energy production that can be associated with this type of
1037 mitigation. Cost and other tradeoffs are important considerations in the implementation of any mitigation
1038 (Best and Halpin, 2019) and the implementation of curtailment can have direct tradeoffs with power
1039 production. As such, we recommend that curtailment and similar turbine operational parameter
1040 approaches be implemented only if bat mortality at offshore wind farms proves to be a key issue. While
1041 there is evidence that migratory tree roosting bats occur offshore, uncertainty remains in both the
1042 abundance and distribution of bats in the offshore environment as well as the degree to which collision
1043 risk and associated mortality mirrors patterns onshore (Solick and Newman, 2021). The further
1044 development and testing of collision detection technologies represents a key step towards better testing of
1045 these types of strategies for the OSW industry.

1046
1047 While curtailment, feathering, and turbine shutdown have been proven effective for bats, at least in
1048 terrestrial settings, additional research is needed for birds before this type of approach is implemented in
1049 the offshore context. There is some evidence that targeted turbine shutdowns may be effective for specific
1050 species (e.g., when they have been detected entering a wind farm), but the greatest challenge to testing
1051 these types of strategies relates to the difficulty in collecting adequate data, including movement data in
1052 real time to inform turbine shutdowns, as well as measuring collisions offshore.

1053 *4.3.4 Minimizing Disturbance*

1054 While many of the mitigation approaches discussed above relate to reducing attraction and/or collision,
1055 behavioral disturbance is a key effect from OSW development for marine birds (Allison et al., 2019; Fox
1056 and Petersen, 2019). In particular, some species avoid offshore wind farms at various scales, which may
1057 result in effective habitat loss or increased energy expenditure (Williams et al., 2024). Birds may also be
1058 disturbed due to vessel and other OSW-related activity, which has the potential to have similar energetic
1059 consequences. Mitigation strategies such as altering vessel parameters to reduce disturbance have the
1060 potential to benefit marine birds, though this has only been tested in a few cases with specific species and
1061 not in the OSW industry specifically. While the impact of this type of disturbance is not well studied, it
1062 may be an important mitigation strategy to evaluate given its potential effectiveness in the maritime
1063 environment. Altering vessel parameters, in particular speed and distance to wildlife, represent key
1064 mitigation approaches used for marine mammals and sea turtles for the OSW and other maritime
1065 industries (Conn and Silber, 2013; Hazel et al., 2007). Thus, additional research to understand variation in
1066 response among species, implementation of best practices, and the potential consequences of vessel
1067 disturbance would go a long way to understand the degree to which this type of effect could be mitigated
1068 effectively for marine birds. There may also be other types of disturbance, such as from aircrafts, which
1069 were not included in our review but may also help reduce disturbance.

1070 **4.4. Importance of Exploring Compensatory Mitigation Measures**

1071 In the absence of proven on-site mitigation approaches (e.g., via avoidance and minimization) to fully
1072 offset effects to aerofauna from OSW development, compensatory mitigation measures should be
1073 explored in the immediate term (Croll et al., 2022). Potential strategies to consider in a compensatory

1074 mitigation framework should include both onsite and offsite approaches and should be considered
1075 regardless of whether the regulatory framework for a given location with OSW development require “no
1076 net loss” of animals or “net gain” of animals in relation to development activities (Edwards-Jones et al.,
1077 2024; U.S. Fish and Wildlife Service (USFWS), 2023). Without on-site mitigation measures that have
1078 been proven to adequately ameliorate effects of development, off-site and potentially out-of-kind
1079 measures must also be considered in order to address remaining unmitigated effects.

1080
1081 There are a range of approaches that have been suggested primarily in the context of other industries. In
1082 the case of birds, most compensatory mitigation has focused on marine birds, and included approaches
1083 related to fisheries (e.g., fisheries closures for species that are key prey for seabirds, efforts to reduce
1084 fisheries bycatch), creation or restoration of nesting habitat and marine reserves, and the eradication and
1085 control of predators, including invasive mammals like rats (Jodice et al., 2019; McGregor et al., 2022). In
1086 general, these types of approaches have required off-site implementation (i.e., away from offshore
1087 facilities) in key foraging habitats and breeding sites; such efforts may require international coordination
1088 to work across different jurisdictions and life history stages. These efforts are often aimed at increasing
1089 survival directly (e.g., reducing mortality via fisheries bycatch) or indirectly (e.g., increasing resource and
1090 habitat availability) or are focused on increasing reproductive success (e.g., reduced predation on nests).
1091 Understanding the effectiveness of these types of approaches is challenging, as it generally requires long-
1092 term monitoring of populations to examine changes in key demographic parameters along with a
1093 quantitative understanding of the influences of other anthropogenic threats on these populations.
1094 Additionally, in some cases there may be no well-established offset approach that is known to be effective
1095 for a given species. However, there is clear evidence that colony-based and island-based mitigation
1096 measures to control invasive predators in the case of seabirds, for example, do increase breeding success
1097 (e.g., (Nogales et al., 2013; Pascal et al., 2008). A recent review focused on OSW development in the
1098 United States noted that compensatory mitigation may be the best approach to adequately prevent
1099 cumulative impacts that lead to population declines across both regional and international scales, and that
1100 implementation will in turn require a strategically coordinated approach (Croll et al., 2022). As with on-
1101 site mitigation, compensatory mitigation measures must be carefully designed and monitored to assess
1102 effectiveness in mitigating OSW impacts.

1103 **4.5 Monitoring Programs to Validate Effectiveness of Mitigation Measures**

1104 Monitoring programs to assess mitigation effectiveness can be difficult to design and implement,
1105 particularly in the offshore context. Studies must be designed with sufficient statistical power, including
1106 acquisition of enough data to reliably detect a change in the selected effect metric if it occurs (Martins et
1107 al., 2023; Regional Synthesis Workgroup of the Environmental Technical Working Group, 2023). This
1108 can be particularly challenging for rare effects, such as collisions, and may require studies of relatively
1109 long duration (e.g., May et al., 2020). These challenges are exacerbated in the offshore environment by
1110 logistical difficulties such as the inability to search for carcasses underneath turbines (Allison et al.,
1111 2019). Monitoring of more common types of effects for OSW, such as displacement, is also complicated
1112 due to high levels of variability in the marine environment, which make it difficult to tease apart effects of
1113 development (or accompanying mitigation measures) from background variation in the distribution of
1114 aerofauna (Lamb et al., 2024; Maclean et al., 2013). Additionally, mitigation approaches implemented
1115 during the project design phase (such as changing turbine height, rotor diameter, air gap, distance between
1116 turbines, turbine layout, etc.) require the comparison of data across multiple wind facilities with differing

1117 characteristics to help assess the effectiveness of these approaches. This is only possible when data
1118 collection methods are harmonized across projects and data are publicly shared, so that multi-facility
1119 studies can be conducted.

1120
1121 A few recent high-profile studies have tested mitigation measures in the field (e.g., (May et al., 2020) or
1122 used meta-analytical frameworks to assess effectiveness of mitigation approaches across multiple studies
1123 (e.g., (Adams et al., 2021; Whitby et al., 2021). A good example of progression in our understanding of
1124 mitigation effectiveness is the approach of painting one turbine blade black in order to reduce motion
1125 smear and increase the visibility of the blades to birds. This approach was first tested in the laboratory in
1126 2003 (Hodos, 2003), and was implemented using a Before-After Control-Impact (BACI) approach at a
1127 terrestrial wind facility in Norway in 2013. Because extensive mortality data was available from the
1128 Norwegian site prior to the treatment (2006-2013), and carcass searches were continued for >3 years after
1129 the treatment was implemented, an 11-year dataset was available to assess the effectiveness of the
1130 treatment. Given the relatively low rate of collisions per year, this long-term dataset was necessary in
1131 order to have sufficient statistical power to ensure that the study could reliably detect a difference
1132 between control and impact scenarios. Indeed, the focal species of study, the White-tailed Eagle
1133 (*Haliaeetus albicilla*), collided with turbines only 7 times throughout this study period (7 prior to
1134 treatment and zero after). There was limited or no evidence of effectiveness for other avian taxa at the site
1135 (May et al., 2020). Several studies are currently underway to replicate the Norway study at other
1136 geographic locations and with other focal species (Vattenfall, 2022). Such replication is essential in order
1137 to assess whether this approach will work in other habitats and for other species with different visual
1138 acuity and behaviors. This example exemplifies the large-scale effort often required to test and validate
1139 mitigation measures.

1140
1141 Given the importance of identifying the most effective mitigation measures to achieve sustainability of
1142 the OSW industry, research to assess the effectiveness of measures is a high priority. There are several
1143 avenues of future research that are important to investigate in the near term. First, we recommend that
1144 research focus on mitigation approaches that have already been implemented at offshore wind farms and
1145 included in project permits in jurisdictions worldwide. This is important since measures that have already
1146 been implemented at existing projects are more likely to also be proposed at new projects. This includes
1147 project design measures such as increasing the air gap to reduce collisions of marine birds. Second, we
1148 recommend focusing research on mitigation approaches with the greatest potential impact in reducing
1149 effects, if proven effective. Curtailment for bats could fall into this category, assuming bats are attracted
1150 to offshore turbines, given the proven effectiveness of this approach onshore (Adams et al., 2021; Whitby
1151 et al., 2021). Curtailment for birds is a commonly discussed measure with little or no evidence of
1152 effectiveness that could also fall into this category. Thirdly, we recommend an immediate focus on
1153 mitigation measures that can be tested at the individual project level, such as visibility enhancement
1154 measures. Studies that require comparisons across projects with different project designs will necessarily
1155 require the construction of multiple projects with differing characteristics and longer-term data collection
1156 and comparisons. In the immediate term, conducting well-designed studies of effects of aerofauna from
1157 OSW development, as well as ensuring the public availability of resulting effects data, is an essential step
1158 to allow for these longer-term studies of the influence of project design on wildlife mitigation.

1159 5. Conclusions

1160 There was no evidence of effectiveness for 86% of the mitigation approaches identified in this literature
1161 review. Of a total of 212 mitigation approaches from 233 source documents, 59% of proposed approaches
1162 were not tested in the reviewed literature to assess whether they were effective at mitigating
1163 anthropogenic impacts to aerofauna. Of the mitigation approaches that were field tested or implemented,
1164 the reviewed literature indicated evidence of their effectiveness in only about 36% of cases. For birds,
1165 approaches related to lighting (reducing artificial light, avoiding white and steady-burning lights, etc.)
1166 were the most commonly tested and effective methods for reducing maladaptive attraction and collisions.
1167 For bats, alteration of turbine operations (e.g., curtailment and feathering of turbine blades) were most
1168 commonly shown to be effective. Avoidance of effects (via careful siting of industrial activity and related
1169 measures to avoid effects to wildlife and their habitats) remains the best available option for mitigation at
1170 OSW facilities.

1171
1172 Minimization and restoration measures were the main focus of this review but there is limited evidence of
1173 effectiveness for most approaches, and we suggest implementation of dedicated testing for commonly
1174 suggested and implemented mitigation measures such as curtailment for birds. Implementation of
1175 unproven mitigations creates the assumption that impacts to aerofauna are being adequately addressed
1176 when that may not be the case. Very few mitigation approaches for birds and bats have sufficient data to
1177 support current implementation at OSW farms without more data being needed on their effectiveness,
1178 either in the short or long term.

1179
1180 Given the current status of available mitigation measures, a sensible approach to aerofauna mitigation at
1181 OSW facilities could include the universal implementation of measures with sufficient evidence of
1182 effectiveness, such as reductions and adjustments to artificial lighting regimes at wind farm structures, as
1183 well as implementation of testing (at multiple locations, focused on a range of species) to further examine
1184 the effectiveness of commonly suggested and implemented mitigation measures for which data are
1185 currently lacking. In the absence of sufficiently validated measures for on-site mitigation of OSW effects
1186 to aerofauna, compensatory mitigation approaches should also be pursued.

1187 6. Glossary

1188 **Aerofauna** - birds and bats

1189
1190 **Air gap** - distance between the surface of the water and the lowest edge of the rotor-swept zone (e.g.,
1191 bottom edge of the turbine blades). Increasing the size of the air gap is a commonly proposed mitigation
1192 measure for some marine bird species, since collision risk models often predict that a larger air gap will
1193 reduce collisions.

1194
1195 **Avoidance** - the first step in the mitigation hierarchy, in which impacts to wildlife or ecosystems are
1196 avoided altogether. Within the context of offshore wind energy development, avoidance of impacts
1197 primarily occurs during planning processes when projects are sited away from high-risk or high-use areas
1198 for wildlife. Note that the term “avoidance” is also used in the offshore wind context to describe the
1199 behavior of aerofauna around turbines or other infrastructure (see also “displacement,” below).

1200 Behavioral avoidance may occur at the macro-, meso- or micro-scale. Animals’ avoidance of individual

1201 turbine blades when they are present within the rotor-swept zone is termed micro-avoidance; meso-
1202 avoidance involves animals present within the broader wind farm footprint staying away from individual
1203 turbines; and avoidance of the entire offshore wind facility is termed macro-avoidance, which can have
1204 resultant changes in wildlife distributions (e.g., displacement).

1205
1206 **Compensation** - the final step in the mitigation hierarchy, in which impacts that cannot be adequately
1207 avoided or minimized are “offset” by activities intended to improve and enhance impacted populations.
1208 Compensation can occur on-site (e.g., where the impact was created) or away from the site of impact, and
1209 may include options such as habitat creation (e.g., new breeding, roosting, wintering sites) and
1210 improvement of existing habitat via invasive species removal or predator control (Arnett and May, 2016).
1211 Compensation aims to ensure that negative environmental impacts that cannot be avoided or minimized
1212 can be moderated by environmental gains to produce net neutral or net positive outcomes for affected
1213 populations and ecosystems.

1214
1215 **Curtailment** - mitigation approach that involves halting operations of wind turbines under specified
1216 circumstances. So-called “blanket” curtailment for bats at terrestrial wind farms typically involves raising
1217 the threshold for ambient wind speed at which turbines begin generating electricity (the turbine “cut-in
1218 speed”), so that turbines are not spinning as quickly at low wind speeds. Curtailment approaches may also
1219 use other environmental conditions and/or detections of animals in the vicinity of the wind farm to inform
1220 curtailment of turbine operations.

1221
1222 **Deterrent** - human invention aimed at scaring or otherwise repulsing wildlife to reduce their presence in
1223 an area. Can be used to reduce potential harm to human interests caused by aeroфаuna (e.g., to reduce crop
1224 damage caused by birds) or to reduce possible harm to wildlife (for example, by attempting to counteract
1225 attraction to a location or structure that may increase collision risk). While some industries distinguish
1226 between deterrents and hazing devices, for the purposes of this review, we use ‘deterrents’ as an
1227 encompassing term to include both. May include acoustic, visual, physical, olfactory and chemical, or
1228 other approaches.

1229
1230 **Displacement** - see “avoidance,” above.

1231
1232 **Effectiveness** - the degree to which a mitigation approach was successful in producing the desired result,
1233 which assumes scientific rigor in data collection for such a conclusion.

1234
1235 **Feathering** - when turbine blades are pitched to catch as little wind as possible, thus slowing the blade
1236 rotation. Below the cut-in wind speed at which turbines begin generating electricity, turbine blades still
1237 spin in the wind but do so much more slowly. In addition to curtailment (above), blades can be feathered
1238 when the turbine is not generating electricity in order to reduce bat mortality at terrestrial wind farms.

1239
1240 **Minimization** - the second step in the mitigation hierarchy, in which expected effects to aeroфаuna are
1241 reduced/minimized to the degree possible. For offshore wind energy development, this primarily occurs
1242 during pre-construction and operations, and can include design decisions (e.g., micro-siting, turbine
1243 height, lighting, addition of perching deterrents) as well as operational parameters (e.g., cut-in speeds at
1244 which turbines begin generating electricity).

1245
1246 **Mitigation** - activities intended to avoid, minimize, or compensate for deleterious anthropogenic impacts
1247 to wildlife or ecosystems. The “mitigation hierarchy” is a commonly accepted framework for conducting
1248 mitigation activities in which avoidance is preferred to minimization, and minimization preferred to
1249 compensation (see definitions above).
1250

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1260 **References**

- 1261 Adams, E.M., Gulka, J., Williams, K.A., 2021. A review of the effectiveness of operational curtailment
1262 for reducing bat fatalities at terrestrial wind farms in North America. PLOS ONE 16, e0256382.
1263 <https://doi.org/10.1371/journal.pone.0256382>
- 1264 Aguilera, E., Knight, R.L., Cummings, J.L., 1991. An Evaluation of Two Hazing Methods for Urban
1265 Canada Geese. Wildl. Soc. Bull. 19, 32–35.
- 1266 Ahlén, I., Bach, L., Baagøe, H.J., Pettersson, J., 2007. Bats and offshore wind turbines studies in southern
1267 Scandinavia. Swedish Environmental Protection Agency, Stockholm, Sweden.
1268 <https://doi.org/10.1644/09-MAMM-S-223R.1>
- 1269 Allison, T.D., Diffendorfer, J.E., Baerwald, E.F., Beston, J.A., Drake, D., Hale, A.M., Hein, C.D., Huso,
1270 M.M., Loss, S.R., Lovich, J.E., Strickland, M.D., Williams, K.A., Winder, V.L., 2019. Impacts to
1271 Wildlife of Wind Energy Siting and Operation in the United States. Issues Ecol.
- 1272 American Wind Wildlife Institute, 2019. AWWI Technical Report: A Summary of Bird Fatality Data in a
1273 Nationwide Database. Washington, DC. Available at www.awwi.org.
- 1274 Arnett, E., Hein, C., Schirmacher, M., Huso, M., Szewczak, J., 2013. Correction: evaluating the
1275 effectiveness of an ultrasonic acoustic deterrent for reducing bat fatalities at wind turbines. PLOS
1276 ONE 8, 10.1371/annotation/a81f59cb-0f82-4c84-a743-895acb4.
- 1277 Arnett, E.B., Brown, W.K., Erickson, W.P., Fiedler, J.K., Hamilton, B.L., Henry, T.H., Jain, A., Johnson,
1278 G.D., Kerns, J., Koford, R.R., Nicholson, C.P., O’Connell, T.J., Piorkowski, M.D., Tankersley,
1279 R.D., 2008. Patterns of bat fatalities at wind energy facilities in North America. J. Wildl. Manag.
1280 72, 61–78. <https://doi.org/10.2193/2007-221>
- 1281 Arnett, E.B., May, R.F., 2016. Mitigating wind energy impacts on wildlife: approaches for multiple taxa.
1282 Hum.-Wildl. Interact. 10, 28–41. <https://doi.org/10.26077/1JEG-7R13>
- 1283 Baasch, D.M., Hegg, A.M., Dwyer, J.F., Caven, A.J., Taddicken, W.E., Worley, C.A., Medaries, A.H.,
1284 Wagner, C.G., Dunbar, P.G., Mittman, N.D., 2022. Mitigating avian collisions with power lines
1285 through illumination with ultraviolet light. Avian Conserv. Ecol. 17.
1286 <https://doi.org/10.5751/ACE-02217-170209>

- 1287 Baerwald, E.F., Edworthy, J., Holder, M., Barclay, R.M.R., 2009. A Large-Scale Mitigation Experiment
1288 to Reduce Bat Fatalities at Wind Energy Facilities. *J. Wildl. Manag.* 73, 1077–1081.
1289 <https://doi.org/10.2193/2008-233>
- 1290 Belant, J.L., Ickes, S.K., 1997. Mylar Flags as Gull Deterrents, in: Proceedings of the Thirteenth Great
1291 Plains Wildlife Damage Control Workshop. pp. 73–80.
- 1292 Bellefleur, D., Lee, P., Ronconi, R.A., 2009. The impact of recreational boat traffic on Marbled Murrelets
1293 (*Brachyramphus marmoratus*). *J. Environ. Manage.* 90, 531–538.
1294 <https://doi.org/10.1016/j.jenvman.2007.12.002>
- 1295 Bennett, V., Hale, A., 2018. Resource Availability May Not Be a Useful Predictor of Migratory Bat
1296 Fatalities or Activity at Wind Turbines. *Diversity* 10, 44. <https://doi.org/10.3390/d10020044>
- 1297 Bennett, V.J., Hale, A.M., 2014. Red aviation lights on wind turbines do not increase bat-turbine
1298 collisions. *Anim. Conserv.* 17, 354–358. <https://doi.org/10.1111/acv.12102>
- 1299 Best, B.D., Halpin, P.N., 2019. Minimizing wildlife impacts for offshore wind energy development:
1300 Winning tradeoffs for seabirds in space and cetaceans in time. *PLOS ONE* 14, e0215722.
1301 <https://doi.org/10.1371/journal.pone.0215722>
- 1302 Blackwell, B.F., Bernhardt, G.E., Dolbeer, R.A., 2002. Lasers at Nonlethal Avian Repellents. *J. Wildl.*
1303 *Manag.* 66, 250–258.
- 1304 Bomford, M., O'Brien, P.H., 1990. Sonic Deterrents in Animal Damage Control: A Review of Device
1305 Tests and Effectiveness. *Wildl. Soc. Bull.* 18, 411–422.
- 1306 Bureau of Ocean Energy Management (BOEM), 2024. Conditions of Construction and Operations Plan
1307 Approval Lease Number OCS-A 0483 [Coastal Virginia Offshore Wind]. U.S. Department of the
1308 Interior, Bureau of Ocean Energy Management, Washington D.C.
- 1309 Bureau of Ocean Energy Management (BOEM), 2023. Conditions of Construction and Operations Plan
1310 Approval Lease Number OCS-A 0486 [Revolution Wind]. U.S. Department of the Interior,
1311 Bureau of Ocean Energy Management, Washington D.C.
- 1312 Bureau of Ocean Energy Management (BOEM), 2022. Conditions of Construction and Operations Plan
1313 Approval Lease Number OCS-A 0517 [South Fork Wind]. U.S. Department of the Interior,
1314 Bureau of Ocean Energy Management, Washington D.C.
- 1315 Bureau of Ocean Energy Management (BOEM), 2021a. Conditions of Construction and Operations Plan
1316 Approval Lease Number OCS-A 0501 [Vineyard Wind 1]. U.S. Department of the Interior,
1317 Bureau of Ocean Energy Management, Washington D.C.
- 1318 Bureau of Ocean Energy Management (BOEM), 2021b. Guidelines for Lighting and Marking of
1319 Structures Supporting Renewable Energy Development. United States Department of the Interior
1320 Bureau of Ocean Energy Management Office of Renewable Energy Programs, Stirling, VA.
- 1321 Burt, C.S., Kelly, J.F., Trankina, G.E., Silva, C.L., Khalighifar, A., Jenkins-Smith, H.C., Fox, A.S.,
1322 Fristrup, K.M., Horton, K.G., 2023. The effects of light pollution on migratory animal behavior.
1323 *Trends Ecol. Evol.*, Special issue: Animal behaviour in a changing world 38, 355–368.
1324 <https://doi.org/10.1016/j.tree.2022.12.006>
- 1325 Christensen-Dalsgaard, S., Dehnhard, N., Moe, B., Systad, G.H.R., Follestad, A., 2019. Unmanned
1326 installations and birds. A desktop study on how to minimize area of conflict (No. NINA Report
1327 173). Norwegian Institute for Nature Research, Trondheim, Norway.
- 1328 Cohen, E.B., Buler, J.J., Horton, K.G., Loss, S.R., Cabrera-Cruz, S.A., Smolinsky, J.A., Marra, P.P.,
1329 2022. Using weather radar to help minimize wind energy impacts on nocturnally migrating birds.
1330 *Conserv. Lett.* 15, e12887. <https://doi.org/10.1111/conl.12887>
- 1331 Conn, P.B., Silber, G.K., 2013. Vessel speed restrictions reduce risk of collision-related mortality for
1332 North Atlantic right whales. *Ecosphere* 4, 1–16. <https://doi.org/10.1890/ES13-00004.1>
- 1333 Conover, M.R., 1984. Comparative Effectiveness of Avitrol, Exploders, and Hawk-Kites in Reducing
1334 Blackbird Damage to Corn. *J. Wildl. Manag.* 48, 109–116.
- 1335 Cook, A.S.C.P., Humphreys, E.M., Bennet, F., Masden, E.A., Burton, N.H.K., 2018. Quantifying avian
1336 avoidance of offshore wind turbines: Current evidence and key knowledge gaps. *Mar. Environ.*
1337 *Res.* 140, 278–288. <https://doi.org/10.1016/j.marenvres.2018.06.017>

- 1338 Cook, A.S.C.P., Ross-Smith, V.H., Roos, S., Burton, N.H.K., Beale, N., Coleman, C., Daniel, H.,
1339 Fitzpatrick, S., Rankin, E., Norman, K., Martin, G., 2011. Identifying a range of options to
1340 prevent or reduce avian collision with offshore wind farms using a UK-based case study. British
1341 Trust for Ornithology, Thetford, UK.
- 1342 Council on Environmental Quality, 2020. Protection of the environment (under the National Environment
1343 Policy Act) (No. Report No 40 CFR 1500-1517). Council on Environmental Quality,
1344 Washington, D.C.
- 1345 Crawford, R.L., Engstrom, R.T., 2001. Characteristics of Avian Mortality at a North Florida Television
1346 tower: a 29-Year Study. *J. Field Ornithol.* 72, 380–388.
- 1347 Croll, D.A., Ellis, A.A., Adams, J., Cook, A.S.C.P., Garthe, S., Goodale, M.W., Hall, C.S., Hazen, E.,
1348 Keitt, B.S., Kelsey, E.C., Leirness, J.B., Lyons, D.E., McKown, M.W., Potiek, A., Searle, K.R.,
1349 Soudijn, F.H., Rockwood, R.C., Tershy, B.R., Tinker, M., VanderWerf, E.A., Williams, K.A.,
1350 Young, L., Zilliagus, K., 2022. Framework for assessing and mitigating the impacts of offshore
1351 wind energy development on marine birds. *Biol. Conserv.* 276, 109795.
1352 <https://doi.org/10.1016/J.BIOCON.2022.109795>
- 1353 Crown Estate Scotland, 2024. Collaboration for Environmental Mitigation & Nature Inclusive Design
1354 (CEMNID) Mitigation Measures Efficacy Review. (No. A-100906-S00-A-REPT-004). Xodus
1355 Group.
- 1356 Cryan, P.M., Brown, A.C., 2007. Migration of bats past a remote island offers clues toward the problem
1357 of bat fatalities at wind turbines. *Biol. Conserv.* 139, 1–11.
- 1358 Cryan, P.M., Gorresen, P.M., Straw, B.R., Thao, S. (Simon), DeGeorge, E., 2022. Influencing Activity of
1359 Bats by Dimly Lighting Wind Turbine Surfaces with Ultraviolet Light. *Animals* 12, 9.
1360 <https://doi.org/10.3390/ani12010009>
- 1361 Curtis, P.D., Smith, C.R., Evans, W., 1993. Techniques for reducing bird use at Nanticoke landfill near
1362 E.A. Link Airport, Broome County, New York, in: Sixth Eastern Wildlife Damage Control
1363 Conference. p. 11.
- 1364 Defingou, M., Bils, F., Horchler, B., Liesenjohann, T., Nehls, G., 2019. PHAROS4MPAs: A Review of
1365 Solutions to Avoid and Mitigate Environmental Impacts of Offshore Windfarms. BioConsult SH
1366 for WWF France, Husum, Germany.
- 1367 Degraer, S., Carey, D.A., Coolen, J.W.P., Hutchison, Z.L., Kerckhof, F., Rumes, B., Vanaverbeke, J.,
1368 2020. Offshore wind farm artificial reefs affect ecosystem structure and functioning: A synthesis.
1369 *Oceanography* 33, 48–57. <https://doi.org/10.5670/oceanog.2020.405>
- 1370 Dias, M.P., Martin, R., Pearmain, E.J., Burfield, I.J., Small, C., Phillips, R.A., Yates, O., Lascelles, B.,
1371 Borboroglu, P.G., Croxall, J.P., 2019. Threats to seabirds: A global assessment. *Biol. Conserv.*
1372 237, 525–537. <https://doi.org/10.1016/J.BIOCON.2019.06.033>
- 1373 Dierschke, V., Furness, R.W., Garthe, S., 2016. Seabirds and offshore wind farms in European waters:
1374 Avoidance and attraction. *Biol. Conserv.* 202, 59–68.
1375 <https://doi.org/10.1016/j.biocon.2016.08.016>
- 1376 Dooling, R.J., 2002. Avian Hearing and the Avoidance of Wind Turbines (No. NREL/TP-500-30844).
1377 National Renewable Energy Laboratory, Golden, CO.
- 1378 Drewitt, A.L., Langston, R.H.W., 2008. Collision effects of wind-power generators and other obstacles on
1379 birds. *Ann. N. Y. Acad. Sci.* 1134, 233–66. <https://doi.org/10.1196/annals.1439.015>
- 1380 Drewitt, A.L., Langston, R.H.W., 2006. Assessing the impacts of wind farms on birds. *Ibis* 148, 29–42.
1381 <https://doi.org/10.1111/j.1474-919X.2006.00516.x>
- 1382 Edwards-Jones, A., Watson, S.C.L., Szostek, C.L., Beaumont, N.J., 2024. Stakeholder insights into
1383 embedding marine net gain for offshore wind farm planning and delivery. *Environ. Chall.* 14,
1384 100814. <https://doi.org/10.1016/j.envc.2023.100814>
- 1385 English, P.A., Mason, T.I., Backstrom, J.T., Tibbles, B.J., Mackay, A.A., Smith, M.J., Mitchell, T., 2017.
1386 Improving Efficiencies of National Environmental Policy Act Documentation for Offshore Wind
1387 Facilities Case Studies Report (No. OCS Study BOEM 2017-026). US Dept. of the Interior,
1388 Bureau of Ocean Energy Management, Sterling, VA.

- 1389 Enos, J.K., Ward, M.P., Hauber, M.E., 2021. A review of the scientific evidence on the impact of
1390 biologically salient frightening devices to protect crops from avian pests. *Crop Prot.* 148, 105734.
1391 <https://doi.org/10.1016/j.cropro.2021.105734>
- 1392 Environment and Climate Change Canada, 2016. Procedures for handling and documenting stranded birds
1393 encountered on infrastructure offshore Atlantic Canada. Environment and Climate Change
1394 Canada, Gatineau, Quebec.
- 1395 Eshleman, K.N., Elmore, A., 2013. Recommended Best Management Practices for Marcellus Shale Gas
1396 Development in Maryland. Maryland Department of the Environment, Baltimore, Maryland.
- 1397 E-TWG Bird and Bat Specialist Committee, 2020. Summary of Discussions from the Bird and Bat
1398 Specialist Committee of the Environmental Technical Working Group (E-TWG). New York
1399 Offshore Wind Environmental Technical Working Group, Albany, NY.
- 1400 Evans, W.R., Akashi, Y., Altman, N.S., Manville, A.M., 2007. Response of night-migrating songbirds in
1401 cloud to colored and flashing light. *North Am. Birds* 60, 476–488.
- 1402 Everaert, J., Stienen, E.W.M., 2007. Impact of wind turbines on birds in Zeebrugge (Belgium):
1403 Significant effect on breeding tern colony due to collisions. *Biodivers. Conserv.* 16, 3345–3359.
1404 <https://doi.org/10.1007/s10531-006-9082-1>
- 1405 Farmer, N.A., Garrison, L.P., Litz, J.A., Ortega-Ortiz, J.G., Rappucci, G., Richards, P.M., Powell, J.R.,
1406 Bethea, D.M., Jossart, J.A., Randall, A.L., Steen, M.E., Matthews, T.N., Morris Jr, J.A., 2023.
1407 Protected species considerations for ocean planning: A case study for offshore wind energy
1408 development in the U.S. Gulf of Mexico. *Mar. Coast. Fish.* 15, e10246.
1409 <https://doi.org/10.1002/mcf2.10246>
- 1410 Fox, A.D., Petersen, I.K., 2019. Offshore wind farms and their effects on birds. *Dan. Orn Foren Tidsskr*
1411 113, 86–101.
- 1412 Frick, W.F., Baerwald, E.F., Pollock, J.F., Barclay, R.M.R., Szymanski, J.A., Weller, T.J., Russell, A.L.,
1413 Loeb, S.C., Medellin, R.A., McGuire, L.P., 2017. Fatalities at wind turbines may threaten
1414 population viability of a migratory bat. *Biol. Conserv.* 209, 172–177.
1415 <https://doi.org/10.1016/j.biocon.2017.02.023>
- 1416 Frick, W.F., Kingston, T., Flanders, J., 2020. A review of the major threats and challenges to global bat
1417 conservation. *Ann. N. Y. Acad. Sci.* 1469, 5–25. <https://doi.org/10.1111/nyas.14045>
- 1418 Gartman, V., Bulling, L., Dahmen, M., Geißler, G., Köppel, J., 2016. Mitigation measures for wildlife in
1419 wind energy development, consolidating the state of knowledge — Part 2: Operation,
1420 decommissioning. *J. Environ. Assess. Policy Manag.* 18, 1650014.
1421 <https://doi.org/10.1142/S1464333216500149>
- 1422 Garvin, J.C., Simonis, J.L., Taylor, J.L., 2024. Does size matter? Investigation of the effect of wind
1423 turbine size on bird and bat mortality. *Biol. Conserv.* 291, 110474.
1424 <https://doi.org/10.1016/j.biocon.2024.110474>
- 1425 Gaultier, S.P., Lilley, T.M., Vesterinen, E.J., Brommer, J.E., 2023. The presence of wind turbines repels
1426 bats in boreal forests. *Landsc. Urban Plan.* 231, 104636.
1427 <https://doi.org/10.1016/j.landurbplan.2022.104636>
- 1428 Gauthreaux, S., Belser, C.G., 2006. Effects of artificial night lighting on migrating birds, in: *Ecological
1429 Consequences of Artificial Night Lighting*. Island Press, Washington, D.C., pp. 67–93.
- 1430 Gehring, J., Kerlinger, P., Manville, A.M., 2009. Communication towers, lights, and birds: successful
1431 methods of reducing the frequency of avian collisions. *Ecol. Appl.* 19, 505–514.
- 1432 Gilmour, L.R.V., Holderied, M.W., Pickering, S.P.C., Jones, G., 2021. Acoustic deterrents influence
1433 foraging activity, flight and echolocation behaviour of free-flying bats. *J. Exp. Biol.* 224,
1434 jeb242715. <https://doi.org/10.1242/jeb.242715>
- 1435 Gilmour, L.R.V., Holderied, M.W., Pickering, S.P.C., Jones, G., 2020. Comparing acoustic and radar
1436 deterrence methods as mitigation measures to reduce human-bat impacts and conservation
1437 conflicts. *PLOS ONE* 15, e0228668. <https://doi.org/10.1371/journal.pone.0228668>
- 1438 Gilsdorf, J.M., Hygnstrom, S.E., VerCauteren, K.C., 2002. Use of frightening devices in wildlife damage
1439 management. *Integr. Pest Manag.* 7, 29–45. <https://doi.org/10.1023/A:1025760032566>

- 1440 Glahn, J.F., Ellis, G., Fioranelli, P., Dorr, B.S., 2000. Evaluation of Moderate and Low-Powered Lasers
1441 for Dispersing Double-crested Cormorants from Their Night Roosts, in: Proceedings of the 9th
1442 Wildlife Damage Management Conference. pp. 34–45.
- 1443 Goodale, M.W., Milman, A., 2020. Assessing cumulative exposure of Northern Gannets to offshore wind
1444 farms. *Wildl. Soc. Bull.* 44, 252–259. <https://doi.org/10.1002/wsb.1087>
- 1445 Goodale, M.W., Milman, A., Griffin, C.R., 2019. Assessing the cumulative adverse effects of offshore
1446 wind energy development on seabird foraging guilds along the East Coast of the United States.
1447 *Environ. Res. Lett.* 14, 074018. <https://doi.org/10.1088/1748-9326/ab205b>
- 1448 Goodale, M.W., Stenhouse, I.J., 2016. A conceptual model to determine vulnerability of wildlife
1449 populations to offshore wind energy development. *Hum.-Wildl. Interact.* 10, 53–61.
1450 <https://doi.org/10.26077/1d31-m472>
- 1451 Guest, E.E., Stamps, B.F., Durish, N.D., Hale, A.M., Hein, C.D., Morton, B.P., Weaver, S.P., Fritts, S.R.,
1452 2022. An Updated Review of Hypotheses Regarding Bat Attraction to Wind Turbines. *Animals*
1453 12. <https://doi.org/10.3390/ani12030343>
- 1454 Hayes, M.A., Lindsay, S.R., Solick, D.I., Newman, C.M., 2023. Simulating the influences of bat
1455 curtailment on power production at wind energy facilities. *Wildl. Soc. Bull.* 47, e1399.
1456 <https://doi.org/10.1002/wsb.1399>
- 1457 Hazel, J., Lawler, I.R., Marsh, H., Robson, S., 2007. Vessel speed increases collision risk for the green
1458 sea turtle *Chelonia mydas*. *Endanger. Species Res.* 3, 105–113.
- 1459 Hein, C., Straw, B., 2021. Proceedings from the State of the Science and Technology for Minimizing
1460 Impacts to Bats from Wind Energy (No. NREL/TP-5000-78557). National Renewable Energy
1461 Laboratory, Golden, CO. <https://doi.org/10.2172/1772994>
- 1462 Hein, C., Williams, K.A., Jenkins, E., 2021. Bat Workgroup Report for the State of the Science Workshop
1463 on Wildlife and Offshore Wind Energy 2020: Cumulative Impacts. New York Energy Research
1464 and Development Authority, Albany, NY. <https://doi.org/10.13140/RG.2.2.24040.60163>
- 1465 Hill, R., Hill, K., Aumuller, R., Schulz, A., Dittmann, T., Kulemeyer, C., Coppack, T., 2014. Of birds,
1466 blades, and barriers: Detecting and analysing mass migration events at alpha ventus, in:
1467 Beiersdorf, A., Wollny-Goerke, K. (Eds.), *Ecological Research at the Offshore Windfarm Alpha*
1468 *Ventus - Challenges, Results, and Perspectives*. Springer Spektrum, Hamburg and Berlin,
1469 Germany, pp. 111–132. <https://doi.org/10.1007/978-3-658-02462-8>
- 1470 Hodos, W., 2003. Minimization of Motion Smear: Reducing Avian Collisions with Wind Turbines (No.
1471 NREL/SR-500-33249). National Renewable Energy Laboratory, Golden, CO.
- 1472 Jägerbrand, A.K., Bouroussis, C.A., 2021. Ecological Impact of Artificial Light at Night: Effective
1473 Strategies and Measures to Deal with Protected Species and Habitats. *Sustainability* 13, 5991.
1474 <https://doi.org/10.3390/su13115991>
- 1475 Jodice, P.G.R., Adams, E.M., Lamb, J., Satgé, Y., Gleason, J.S., 2019. GoMAMN Strategic Bird
1476 Monitoring Guidelines: Seabirds, in: *Strategic Bird Monitoring Guidelines for the Northern Gulf*
1477 *of Mexico*. Mississippi Agricultural and Forestry Experiment Station Research Bulletin,
1478 Mississippi State University, pp. 129–170.
- 1479 Johnson, G.D., Erickson, W.P., Strickland, M.D., Shepherd, M.F., Shepherd, D.A., Sarappo, S.A., 2003.
1480 Mortality of Bats at a Large-scale Wind Power Development at Buffalo Ridge, Minnesota. *Am.*
1481 *Midl. Nat.* 150, 332–342. [https://doi.org/10.1674/0003-](https://doi.org/10.1674/0003-0031(2003)150[0332:MOBAAL]2.0.CO;2)
1482 [0031\(2003\)150\[0332:MOBAAL\]2.0.CO;2](https://doi.org/10.1674/0003-0031(2003)150[0332:MOBAAL]2.0.CO;2)
- 1483 Kerlinger, P., Gehring, J.L., Erickson, W.P., Curry, R., Jain, A., Guarnaccia, J., 2010. Night migrant
1484 fatalities and obstruction lighting at wind turbines in North America. *Wilson J. Ornithol.* 122,
1485 744–754. <https://doi.org/10.1676/06-075.1>
- 1486 Kiesecker, J.M., Copeland, H., Pocewicz, A., McKenney, B., 2010. Development by design: blending
1487 landscape-level planning with the mitigation hierarchy. *Front. Ecol. Environ.* 8, 261–266.
1488 <https://doi.org/10.1890/090005>
- 1489 Lagerveld, S., Wilkes, T., van Puijenbroek, M.E.B., Noort, B.C.A., Geelhoed, S.C.V., 2023. Acoustic
1490 monitoring reveals spatiotemporal occurrence of *Nathusius' pipistrelle* at the southern North Sea

- 1491 during autumn migration. *Environ. Monit. Assess.* 195, 1016. <https://doi.org/10.1007/s10661->
1492 023-11590-2
- 1493 Lamb, J., Gulka, J., Adams, E., Cook, A., Williams, K.A., 2024. A synthetic analysis of post-construction
1494 displacement and attraction of marine birds at offshore wind energy installations. *Environ. Impact*
1495 *Assess. Rev.* 108, 107611. <https://doi.org/10.1016/j.eiar.2024.107611>
- 1496 Langston, R., Pullan, J., 2003. *Windfarms and Birds: An analysis of the effects of windfarms on birds,*
1497 *and guidance on environmental assessment criteria and site selection issues (No. Report T-*
1498 *PVS/Inf).* BirdLife International to the Council of Europe, Bern Convention on the Conservation
1499 of European Wildlife and Natural Habitats, Strasbourg, Germany.
- 1500 Langston, R.H.W., 2013. Birds and wind projects across the pond: A UK perspective. *Wildl. Soc. Bull.*
1501 37, 5–18. <https://doi.org/10.1002/wsb.262>
- 1502 Liang, X., Yang, C., Zhang, Y., Xue, Y., 2023. Mitigating the Negative Impact of Wind Power on
1503 Soaring Birds through Government Restrictions. *Energies* 16, 6584.
1504 <https://doi.org/10.3390/en16186584>
- 1505 Long, C.V., Flint, J.A., Lepper, P.A., 2011. Insect attraction to wind turbines: Does colour play a role?
1506 *Eur. J. Wildl. Res.* 57, 323–331. <https://doi.org/10.1007/s10344-010-0432-7>
- 1507 Lüdeke, J., 2019. Retrospective and Outlook for an Environmentally Sound Development of Offshore
1508 Wind Energy. *Int. J. Environ. Sci.* 4, 1–19.
- 1509 Maclean, I.M.D., Rehfish, M.M., Skov, H., Thaxter, C.B., 2013. Evaluating the statistical power of
1510 detecting changes in the abundance of seabirds at sea. *Ibis* 155, 113–126.
1511 <https://doi.org/10.1111/j.1474-919X.2012.01272.x>
- 1512 Madsen, P., Wahlberg, M., Tougaard, J., Lucke, K., Tyack, P., 2006. Wind turbine underwater noise and
1513 marine mammals: implications of current knowledge and data needs. *Mar. Ecol. Prog. Ser.* 309,
1514 279–295. <https://doi.org/10.3354/meps309279>
- 1515 Marques, A.T., Batalha, H., Rodrigues, S., Costa, H., João, M., Pereira, R., Fonseca, C., Mascarenhas,
1516 M., Bernardino, J., 2014. Understanding bird collisions at wind farms: An updated review on the
1517 causes and possible mitigation strategies. *Biol. Conserv.* 179, 40–52.
1518 <https://doi.org/10.1016/j.biocon.2014.08.017>
- 1519 Martins, R.C., Bernardino, J., Moreira, F., 2023. A review of post-construction monitoring practices used
1520 in the evaluation of transmission power line impacts on birds and mitigation effectiveness, with
1521 proposals for guideline improvement. *Environ. Impact Assess. Rev.* 100, 107068.
1522 <https://doi.org/10.1016/j.eiar.2023.107068>
- 1523 May, R., Nygård, T., Falkdalen, U., Åström, J., Hamre, Ø., Stokke, B.G., 2020. Paint it black: Efficacy of
1524 increased wind turbine rotor blade visibility to reduce avian fatalities. *Ecol. Evol.* 10, 8927–8935.
1525 <https://doi.org/10.1002/ece3.6592>
- 1526 May, R., Reitan, O., Bevanger, K., Lorentsen, S.-H., Nygård, T., 2015. Mitigating wind-turbine induced
1527 avian mortality: Sensory, aerodynamic and cognitive constraints and options. *Renew. Sustain.*
1528 *Energy Rev.* 42, 170–181. <https://doi.org/10.1016/j.rser.2014.10.002>
- 1529 McGregor, R., Trinder, M., Goodship, N., 2022. Assessment of compensatory measures for impacts of
1530 offshore windfarms on seabirds (No. Report NECR431). Natural England, York, UK.
- 1531 Mockrin, M.H., Gravenmier, R.A., 2012. Synthesis of wind energy development and potential impacts on
1532 wildlife in the Pacific Northwest, Oregon and Washington. United States Department of
1533 Agriculture, Forest Service, Pacific Northwest Research Station, Portland, OR.
- 1534 Montevecchi, W.A., 2006. Influences of artificial light on marine birds, in: Rich, C., Longcore, T. (Eds.),
1535 *Ecological Consequences of Artificial Night Lighting.* Island Press, Washington, D.C., pp. 94–
1536 113. <https://doi.org/10.1111/bph.13539>
- 1537 Murgatroyd, M., Bouten, W., Amar, A., 2021. A predictive model for improving placement of wind
1538 turbines to minimise collision risk potential for a large soaring raptor. *J. Appl. Ecol.* 58, 857–868.
1539 <https://doi.org/10.1111/1365-2664.13799>

- 1540 Neate-Clegg, M.H.C., Horns, J.J., Adler, F.R., Kemahlı Aytekin, M.Ç., Şekercioğlu, Ç.H., 2020.
1541 Monitoring the world's bird populations with community science data. *Biol. Conserv.* 248,
1542 108653. <https://doi.org/10.1016/j.biocon.2020.108653>
- 1543 Nicholls, B., Racey, P.A., 2009. The aversive effect of electromagnetic radiation on foraging bats - A
1544 possible means of discouraging bats from approaching wind turbines. *PLoS ONE* 4, e6246.
1545 <https://doi.org/10.1371/journal.pone.0006246>
- 1546 Nicholls, B., Racey, P.A., 2007. Bats Avoid Radar Installations: Could Electromagnetic Fields Deter Bats
1547 From Colliding With Wind Turbines? *PLoS ONE* 2, e297.
1548 <https://doi.org/10.1371/journal.pone.0000297>
- 1549 Nogales, M., Vidal, E., Medina, F.M., Bonnaud, E., Tershy, B.R., Campbell, K.J., Zavaleta, E.S., 2013.
1550 Feral Cats and Biodiversity Conservation: The Urgent Prioritization of Island Management.
1551 *BioScience* 63, 804–810. <https://doi.org/10.1525/bio.2013.63.10.7>
- 1552 ONeil, D.R., 2020. Reducing Bat Fatalities Using Ultrasonic Acoustic Deterrent Technology: A Potential
1553 Mechanism for Conservation at Offshore Wind Energy Sites (M.S. thesis). Harvard Extension
1554 School, Boston, MA.
- 1555 Pascal, M., Lorvelec, O., Bretagnolle, V., Culioli, J., 2008. Improving the breeding success of a colonial
1556 seabird: a cost-benefit comparison of the eradication and control of its rat predator. *Endanger.*
1557 *Species Res.* 4, 267–276. <https://doi.org/10.3354/esr00080>
- 1558 Peste, F., Paula, A., da Silva, L.P., Bernardino, J., Pereira, P., Mascarenhas, M., Costa, H., Vieira, J.,
1559 Bastos, C., Fonseca, C., Pereira, M.J.R., 2015. How to mitigate impacts of wind farms on bats? A
1560 review of potential conservation measures in the European context. *Environ. Impact Assess. Rev.*
1561 51, 10–22. <https://doi.org/10.1016/j.eiar.2014.11.001>
- 1562 Poot, H., Ens, B.J., Vries, H. de, Donners, M.A.H., Wernand, M.R., Marquenie, J.M., 2008. Green light
1563 for nocturnally migrating birds. *Ecol. Soc.* 13, 47.
- 1564 Randall, A.L., Jossart, J.A., Matthews, T., Steen, M., Boube, I., Stradley, S., Del Rio, R., Inzinna, D.,
1565 Oos, C., Coats, L., Shin, G., Griffith, C., Morris Jr, J.A., 2022. A Wind Energy Area Siting
1566 Analysis for the Gulf of Mexico Call Area. National Oceanic and Atmospheric Administration,
1567 Bureau of Ocean Energy Management, Beaufort, NC; New Orleans, LA.
- 1568 Rebke, M., Dierschke, V., Weiner, C.N., Aumüller, R., Hill, K., Hill, R., 2019. Attraction of nocturnally
1569 migrating birds to artificial light: The influence of colour, intensity and blinking mode under
1570 different cloud cover conditions. *Biol. Conserv.* 233, 220–227.
1571 <https://doi.org/10.1016/J.BIOCON.2019.02.029>
- 1572 Regional Synthesis Workgroup of the Environmental Technical Working Group, 2023. Responsible
1573 Practices for Regional Wildlife Monitoring and Research in Relation to Offshore Wind Energy
1574 Development. New York Offshore Wind Environmental Technical Working Group, Albany, NY.
- 1575 Reusch, C., Lozar, M., Kramer-Schadt, S., Voigt, C.C., 2022. Coastal onshore wind turbines lead to
1576 habitat loss for bats in Northern Germany. *J. Environ. Manage.* 310, 114715.
1577 <https://doi.org/10.1016/j.jenvman.2022.114715>
- 1578 Ronconi, R.A., St. Clair, C.C., 2002. Management Options to Reduce Boat Disturbance on Foraging
1579 Black Guillemots (*Cepphus grylle*) in the Bay of Fundy. *Biol. Conserv.* 108, 265–271.
1580 [https://doi.org/10.1016/S0006-3207\(02\)00126-X](https://doi.org/10.1016/S0006-3207(02)00126-X)
- 1581 Russell, D.J.F., Brasseur, S.M.J.M., Thompson, D., Hastie, G.D., Janik, V.M., Aarts, G., McClintock,
1582 B.T., Matthiopoulos, J., Moss, S.E.W., McConnell, B., 2014. Marine mammals trace
1583 anthropogenic structures at sea. *Curr. Biol.* 24, 638–639.
1584 <https://doi.org/10.1016/j.cub.2014.06.033>
- 1585 Rydell, J., Engström, H., Hedenström, A., Larsen, J.K., Pettersson, J., Green, M., 2012. The effect of
1586 wind power on birds and bats-A synthesis (No. Report 6511). Swedish Environmental Protection
1587 Agency, Stockholm, Sweden. <https://doi.org/10.1007/s10726-011-9277-z>
- 1588 Schirmacher, M.R., 2020. Evaluating the Effectiveness of an Ultrasonic Acoustic Deterrent in Reducing
1589 Bat Fatalities at Wind Energy Facilities (No. DOE-BCI-0007036). Bat Conservation
1590 International, Austin, TX. <https://doi.org/10.2172/1605929>

- 1591 Schuster, E., Bulling, L., Köppel, J., 2015. Consolidating the State of Knowledge: A Synoptical Review
1592 of Wind Energy's Wildlife Effects. *Environ. Manage.* 56, 300–331.
1593 <https://doi.org/10.1007/s00267-015-0501-5>
- 1594 Schwemmer, P., Mendel, B., Sonntag, N., Dierschke, V., Garthe, S., 2011. Effects of ship traffic on
1595 seabirds in offshore waters: Implications for marine conservation and spatial planning. *Ecol.*
1596 *Appl.* 21, 1851–1860. <https://doi.org/10.1890/10-0615.1>
- 1597 Sherman, D.E., Barras, A.E., 2004. Efficacy of a Laser Device for Hazing Canada Geese from Urban
1598 Areas of Northeast Ohio. *Ohio J. Sci.* 104, 38–42.
- 1599 Simmons, R.E., Ralston-Paton, S., Colyn, R., Garcia-Heras, M.-S., 2020. Black Harriers and Wind
1600 Energy: Guidelines for impact assessment, monitoring and mitigation. *BirdLife South Africa*,
1601 Johannesburg, South Africa.
- 1602 Sinclair, K., DeGeorge, E., 2016. Framework for Testing the Effectiveness of Bat and Eagle Impact-
1603 Reduction Strategies at Wind Energy Projects. National Renewable Energy Laboratory, U.S.
1604 Department of Energy, Golden, CO.
- 1605 Skov, H., Heinanen, S., Norman, T., Ward, R.M., Mendez-Roldan, S., Ellis, I., 2018. ORJIP Bird
1606 Collision and Avoidance Study. Final Report - April 2018. The Carbon Trust, United Kingdom.
- 1607 Smallwood, K.S., Bell, D.A., 2020. Effects of Wind Turbine Curtailment on Bird and Bat Fatalities. *J.*
1608 *Wildl. Manag.* <https://doi.org/10.1002/jwmg.21844>
- 1609 Solick, D.I., Newman, C.M., 2021. Oceanic records of North American bats and implications for offshore
1610 wind energy development in the United States. *Ecol. Evol.* 11, 14433–14447.
1611 <https://doi.org/10.1002/ece3.8175>
- 1612 Stantec Consulting Services Inc., 2016. Long-term bat monitoring on islands, offshore structures, and
1613 coastal sites in the Gulf of Maine, mid-Atlantic, and Great Lakes - Final Report. U.S. Department
1614 of Energy, Tompsham, ME.
- 1615 Stickley, A.R., King, J.O., 1993. Long-Term Trial of An Inflatable Effigy Scare Device for Repelling
1616 Cormorants from Catfish Ponds, in: *Proceedings of the 6th Eastern Wildlife Damage Control*
1617 *Conference*. pp. 1–5.
- 1618 Stone, E.L., Harris, S., Jones, G., 2015. Impacts of artificial lighting on bats: A review of challenges and
1619 solutions. *Mamm. Biol.* 80, 213–219. <https://doi.org/10.1016/j.mambio.2015.02.004>
- 1620 Syposz, M., Padget, O., Willis, J., Doren, B.M.V., Gillies, N., Fayet, A.L., Wood, M.J., Alejo, A.,
1621 Guilford, T., 2021. Avoidance of different durations, colours and intensities of artificial light by
1622 adult seabirds. *Sci. Rep.* 1–13. <https://doi.org/10.1038/s41598-021-97986-x>
- 1623 Thaxter, C.B., Ross-Smith, V.H., Bouten, W., Clark, N.A., Conway, G.J., Masden, E.A., Clewley, G.D.,
1624 Barber, L.J., Burton, N.H.K., 2019. Avian vulnerability to wind farm collision through the year:
1625 Insights from lesser black-backed gulls (*Larus fuscus*) tracked from multiple breeding colonies. *J.*
1626 *Appl. Ecol.* 56, 2410–2422. <https://doi.org/10.1111/1365-2664.13488>
- 1627 Thelander, C.G., Smallwood, K.S., Ruge, L., 2003. Bird Risk Behaviors and Fatalities at the Altamont
1628 Pass Wind Resource Area: Period of Performance, March 1998--December 2000 (No. NREL/SR-
1629 500-33829, 15006013). National Renewable Energy Laboratory, Golden, CO.
1630 <https://doi.org/10.2172/15006013>
- 1631 Thomsen, F., Lüdemann, K., Kafemann, R., Piper, W., 2006. Effects of offshore wind farm noise on
1632 marine mammals and fish. COWRIE Ltd., Hamburg, Germany.
- 1633 Thurber, B.G., Kilpatrick, R.J., Tang, G.H., Wakim, C., Zimmerling, J.R., 2023. Economic Impacts of
1634 Curtailing Wind Turbine Operations for the Protection of Bat Populations in Ontario. *Wind* 3,
1635 291–301. <https://doi.org/10.3390/wind3030017>
- 1636 Tjørnløv, R.S., Skov, H., Armitage, M., Barker, M., Jørgensen, J.B., Mortensen, L.O., Thomas, K.,
1637 Uhrenholdt, T., 11820296, 2023. Resolving Key Uncertainties of Seabird Flight and Avoidance
1638 Behaviours at Offshore Wind Farms: Final report for the study period 2020-2021 (No. Project
1639 No. 11820296). Vattenfall, Horsholm, Denmark.

- 1640 Tulp, I., Schekkerman, H., Larsen, J.K., Van Der Winden, J., Van De Haterd, R.J.W., Van Horssen, P.,
1641 Dirksen, S., Spaans, A.L., 1999. Nocturnal Flight Activity of Sea Ducks Near the Windfarm Tunø
1642 Knob in the Kattegat (No. report nr. 99.64). Bureau Waardenburg, Culemborg, Netherlands.
1643 U.S. Fish and Wildlife Service (USFWS), 2023. U.S. Fish & Wildlife Service Mitigation Policy
1644 (Appendix 1, 501 FW 2). U.S. Fish and Wildlife Service, Arlington, VA.
1645 U.S. Fish and Wildlife Service (USFWS), 2021. Recommended Best Practices for Communication Tower
1646 Design, Siting, Construction, Operation, Maintenance, and Decommissioning. U.S. Fish and
1647 Wildlife Service, Migratory Bird Program, Falls Church, VA.
1648 U.S. Fish and Wildlife Service (USFWS), 2012. U.S. Fish and Wildlife Service Land-based Wind Energy
1649 Guidelines (No. OMB Control No 1018-0148). U.S. Fish and Wildlife Service, Arlington, VA.
1650 Vaissière, A.-C., Levreil, H., Pioch, S., Carlier, A., 2014. Biodiversity Offsets for Offshore Wind Farm
1651 Projects: The Current Situation in Europe. *Mar. Policy* 48, 172–183.
1652 <https://doi.org/10.1016/j.marpol.2014.03.023>
1653 van de Laar, I.F.J.T., 2007. Green light to Birds: Investigation Into the Effect of Bird-friendly Lighting
1654 (No. NAM Locatie L15-FA-1). Nederlandse Aardolie Maatschappij, Assen, Netherlands.
1655 Van Hoorick, G., 2014. Compensatory Measures in European Nature Conservation Law. *Utrecht Law*
1656 *Rev.* 10, 161. <https://doi.org/10.18352/ulr.276>
1657 Vattenfall, 2022. Research into the effect of black blade in wind turbine. URL
1658 [https://group.vattenfall.com/press-and-media/newsroom/2022/black-turbine-blades-reduce-bird-](https://group.vattenfall.com/press-and-media/newsroom/2022/black-turbine-blades-reduce-bird-collisions)
1659 [collisions](https://group.vattenfall.com/press-and-media/newsroom/2022/black-turbine-blades-reduce-bird-collisions) (accessed 8.12.24).
1660 Verfuss, U.K., Sparling, C.E., Arnot, C., Judd, A., Coyle, M., 2016. Review of Offshore Wind Farm
1661 Impact Monitoring and Mitigation with Regard to Marine Mammals, in: Popper, A.N., Hawkins,
1662 A. (Eds.), *The Effects of Noise on Aquatic Life II, Advances in Experimental Medicine and*
1663 *Biology*. Springer New York, New York, NY, pp. 1175–1182. [https://doi.org/10.1007/978-1-](https://doi.org/10.1007/978-1-4939-2981-8_147)
1664 [4939-2981-8_147](https://doi.org/10.1007/978-1-4939-2981-8_147)
1665 Voigt, C.C., Rehnig, K., Lindecke, O., Pētersons, G., 2018. Migratory bats are attracted by red light but
1666 not by warm-white light: Implications for the protection of nocturnal migrants. *Ecol. Evol.* 1–9.
1667 <https://doi.org/10.1002/ece3.4400>
1668 Walsh, C., Hüpopp, O., Karwinkel, T., Liedvogel, M., Lindecke, O., McLaren, J.D., Schmaljohann, H.,
1669 Siebenhüner, B., 2024. Light Pollution at Sea: Implications and Potential Hazards of Human
1670 Activity for Offshore Bird and Bat Movements in the Greater North Sea.
1671 <https://doi.org/10.32942/X2P60W>
1672 Watson, R.T., Kolar, P.S., Ferrer, M., Nygård, T., Johnston, N., Hunt, W.G., Smit-Robinson, H.A.,
1673 Farmer, C.J., Huso, M., Katzner, T.E., 2018. Raptor Interactions With Wind Energy: Case
1674 Studies From Around the World. *J. Raptor Res.* 52, 1–18. <https://doi.org/10.3356/JRR-16-100.1>
1675 Werber, Y., Hareli, G., Yinon, O., Sapir, N., Yovel, Y., 2023. Drone-mounted audio-visual deterrence of
1676 bats: implications for reducing aerial wildlife mortality by wind turbines. *Remote Sens. Ecol.*
1677 *Conserv.* 9, 404–419. <https://doi.org/10.1002/rse2.316>
1678 Whisson, D.A., Takekawa, J.Y., 2000. Testing the Effectiveness of an Aquatic Hazing Device on
1679 Waterbirds in the San Francisco Bay Estuary of California. *Waterbirds Int. J. Waterbird Biol.* 23,
1680 56–63.
1681 Whitby, M.D., Schirmacher, M.R., Frick, W.F., 2021. The State of the Science on Operational
1682 Minimization to Reduce Bat Fatality at Wind Energy Facilities., A report submitted to the
1683 National Renewable Energy Laboratory. Bat Conservation International, Austin, Texas.
1684 Wiese, F.K., Montevecchi, W.A., Davoren, G.K., Huettmann, F., Diamond, A.W., Linke, J., 2001.
1685 Seabirds at risk around offshore oil platforms in the Northwest Atlantic. *Mar. Pollut. Bull.* 42,
1686 1285–1290. [https://doi.org/10.1016/S0025-326X\(01\)00096-0](https://doi.org/10.1016/S0025-326X(01)00096-0)
1687 Williams, K., Connelly, E., Johnson, S., Stenhouse, I., 2015. Wildlife Densities and Habitat Use Across
1688 Temporal and Spatial Scales on the Mid-Atlantic Outer Continental Shelf: Final Report to the
1689 Department of Energy EERE Wind & Water Power Technologies Office, Award Number: DE-
1690 EE0005362 (No. BRI 2015-11). Biodiversity Research Institute, Portland, ME.

- 1691 Williams, K.A., Gulka, J., Cook, A.S.C.P., Diehl, R.H., Farnsworth, A., Goyert, H., Hein, C., Loring, P.,
1692 Mizrahi, D., Petersen, I.K., Peterson, T., Press, K.M., Stenhouse, I.J., 2024. A framework for
1693 studying the effects of offshore wind energy development on birds and bats in the Eastern United
1694 States. *Front. Mar. Sci.* 11. <https://doi.org/10.3389/fmars.2024.1274052>
1695 World Bank Group, 2015. Environmental, Health, and Safety Guidelines for Wind Energy. The World
1696 Bank Group. <https://doi.org/10.1016/j.jpowsour.2006.07.041>
1697